

## **Life in a drop: sampling environmental DNA for marine fishery management and ecosystem monitoring**

### **Abstract**

Science-based management of marine fisheries and effective ecosystem monitoring both require the analysis of large amounts of often complex and difficult to collect information. Legislation also increasingly requires the attainment of good environmental status, which again demands collection of data to enable efficient monitoring and management of biodiversity. Such data is traditionally obtained as a result of research surveys through the capture and/or visual identification of organisms. Recent years have seen significant advances in the utilisation of environmental DNA (eDNA) in the marine environment in order to develop alternative cost-effective ways to gather relevant data. Such approaches attempt to identify and/or quantify the species present at a location through the detection of extra-organismal DNA in the environment. These new eDNA based approaches have the potential to revolutionise data collection in the marine environment using non-invasive sampling methods and providing snapshots of biodiversity beyond the capacity of traditional sampling. Here we present a non-technical summary of different approaches in the field of eDNA, and emphasise the broad application of this approach, with value for the governance and management of marine aquatic ecosystems. The review focuses on identifying those tools which are now readily applicable and those which show promise but are currently in development and require further validations. The aim is to provide an understanding of techniques and concepts that can be used by managers without genetic or genomic expertise when consulting with specialists to perform joint evaluations of the utility of the approaches.

## Life in a drop: sampling environmental DNA for marine fishery management and ecosystem monitoring

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## 3 Abstract

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6 increasingly requires the attainment of good environmental status, which again demands collection  
7 of data to enable efficient monitoring and management of biodiversity. Such data is traditionally  
8 obtained as a result of research surveys through the capture and/or visual identification of organisms.  
9 Recent years have seen significant advances in the utilisation of environmental DNA (eDNA) in the  
10 marine environment in order to develop alternative cost-effective ways to gather relevant data. Such  
11 approaches attempt to identify and/or quantify the species present at a location through the  
12 detection of extra-organismal DNA in the environment. These new eDNA based approaches have the  
13 potential to revolutionise data collection in the marine environment using non-invasive sampling  
14 methods and providing snapshots of biodiversity beyond the capacity of traditional sampling. Here we  
15 present a non-technical summary of different approaches in the field of eDNA, and emphasise the  
16 broad application of this approach, with value for the governance and management of marine aquatic  
17 ecosystems. The review focuses on identifying those tools which are now readily applicable and those  
18 which show promise but are currently in development and require further validations. The aim is to  
19 provide an understanding of techniques and concepts that can be used by managers without genetic  
20 or genomic expertise when consulting with specialists to perform joint evaluations of the utility of the  
21 approaches.

22  
23 **Keywords:** Environmental DNA, eDNA, management, ecosystem, fisheries

25

26 **1. Introduction**

27 Globally, it is increasingly acknowledged that our future depends on the maintenance of good  
28 environmental status and the conservation of biodiversity, both within defined regional and global  
29 standards [1, 2]. The broad consensus is endorsed by such global initiatives as the UN Sustainable  
30 Development Goals [3]. Moreover, international and national policies and legislation require the  
31 protection of the environment and ecosystems [4-6]. For example, this is explicitly aimed at under the  
32 remit of the development of an international instrument on marine biodiversity in areas beyond  
33 national jurisdiction (ABNJ) and stipulated in the European Union Marine Strategy Framework  
34 Directive [7], and also the Common Fisheries Policy (CFP). The implementation of such legal  
35 requirements requires commitment of the member states to carry out extensive monitoring in time  
36 and space, preferably in real-time. The development of tools to assess impacts such as invasive species  
37 introduction and spread, climate change, contaminants, eutrophication, fishing activities and marine  
38 litter on populations and ecosystem interactions remains a high priority. This is an increasingly  
39 challenging undertaking, to which state-of-the-art technological and scientific developments can and  
40 should contribute.

41 Effective ecosystem monitoring, the sustainable exploitation of aquatic living resources,  
42 sustainable fisheries management and associated policy development should be, as in the case of the  
43 CFP, a legally enshrined requirement, based on the best available scientific advice. The integration of  
44 scientific advice into governance and policy development and implementation is often challenging,  
45 particularly the communication of scientific approaches from specialists to managers and policy  
46 makers in a rapidly developing and specialised field. This review seeks to address this issue with  
47 regards to new genetic based techniques in the fields of species identification and community  
48 characterisation and thus facilitate more effective development of marine fishery management and  
49 monitoring approaches.

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50 Effective fishery and ecosystem management rely on the identification and quantification of the  
51 species living a certain environment, that is, characterising its biodiversity. There are two significant  
52 limitations in gathering such information using traditional techniques: how to representatively sample  
53 the biodiversity in an ecosystem and how to identify individuals to species level? Sampling requires  
54 complicated logistics, is costly, is biased in its sampling coverage, and is especially difficult for species  
55 with low abundance and/or elusive species. Identification also requires taxonomic expertise, which is  
56 often lacking and difficult to apply in some cryptic species. The requirement to overcome such  
57 impediments has stimulated the search for new tools and approaches to integrate the various  
58 environmental dimensions in decision making into an evidence-based policy approach [8]. One such  
59 approach is utilisation of DNA collected from the environment to identify and/or quantify the species  
60 present in the ecosystem.

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61 Environmental DNA (eDNA) stems from individual organisms which release DNA into the  
62 environment through waste products, skin/tissue, scales, gametes, mucus, blood and carcasses [9-  
63 12]. This extra-organismal DNA is termed environmental DNA (eDNA) [13]. In contrast to DNA  
64 extracted from tissue samples, or community DNA – where DNA is extracted from communities of  
65 whole organisms - eDNA does not require sampling the target organisms themselves, but instead the  
66 sampling of the environment they live in [14, 15]. The development of new ways of monitoring marine  
67 ecosystems and marine biodiversity using eDNA has advanced over recent years and has  
68 revolutionised the ability to track invasive species, monitor endangered species, assess the health of  
69 fish stocks, and explore the world of marine biodiversity [16]. The seeming simplicity and cost-  
70 effectiveness of eDNA-based approaches, together with the interest from wider stakeholder groups,  
71 has made such applications highly attractive [17].

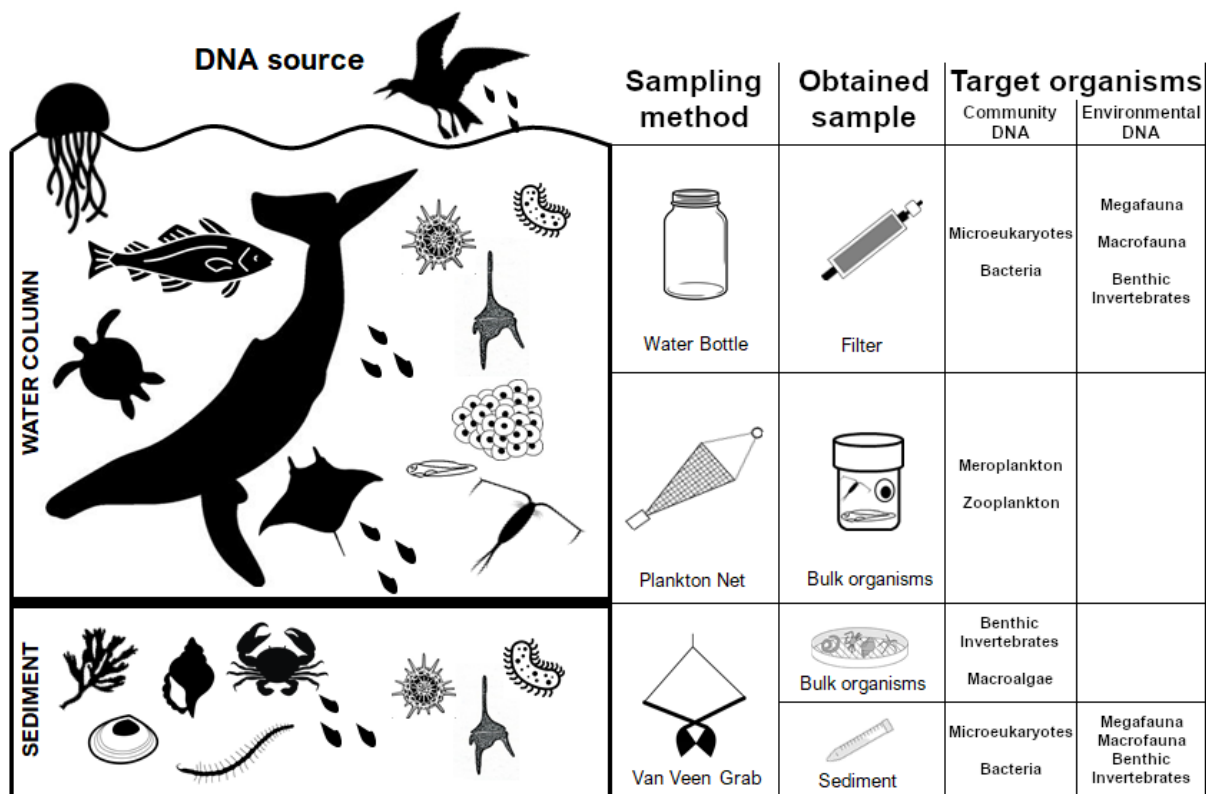
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72 The development of genetic technologies to identify species and characterise whole  
73 communities through the collection and filtration of water and/or sediment sample is both a  
74 potentially invaluable tool for managers and an irresistible story for the popular press. Press articles  
75 focusing on such tools range from the very small, such as “New Nano Strategy Fights Superbugs” [18],

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76 to the very large (and improbable) “Loch Ness Monster Hunters to Try DNA Search?” [19].  
77 Disentangling fact from fiction, and hyperbola from reality, is thus not a simple task for the manager  
78 striving to understand the field. As such this raises two opposing issues which could each negatively  
79 affect the ability to manage fisheries and monitor ecosystems using the most appropriate available  
80 scientific tools: the pre-emptive uptake of unproven approaches versus the failure to take advantage  
81 of robust new techniques. Stories in the press, together with questions from stakeholders, about new  
82 potential approaches that have been developed are often powerful incentives for major funding and  
83 uptake of these tools in practice [20]. Whilst in some cases this uptake may be justified, in others,  
84 especially in rapidly developing fields, such reliance may be potentially premature. However, each  
85 investment requires an accessible, robust and balanced evidence base as deriving management  
86 decisions on unproven and/or unreliable techniques brings obvious dangers and potential lack of trust  
87 in novel molecular technologies. Further, focusing effort and especially funding on such approaches  
88 means that other, perhaps more proven techniques with higher TRL (technology readiness levels) will  
89 be starved of resources. It is thus of particular importance that managers and policy makers can  
90 distinguish with confidence among approaches that although show promise, are at an early stage of  
91 validation.

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92 The converse of the dangers of using unproven tools is avoiding the utilisation of effective  
93 proven tools due to uncertainties about their efficacy. As scientific technologies develop it is often the  
94 case that some areas progress further and faster than others. Proven approaches emerge and begin  
95 to be utilised in limited applications. In order to take full advantage of such developments in a wider  
96 context, managers need a straightforward guideline explaining the potential of each molecular tool  
97 and its state of readiness for routine applications in order to navigate in the various information  
98 streams and stakeholder drivers they are exposed to.

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99 In order to bridge the information gap between the specialist and the manager, we provide here  
100 a non-technical synthesis of the evidence surrounding the use of eDNA based monitoring techniques  
101 for management of fisheries and ecosystems in the marine environment. It is not intended to be an

102 exhaustive overview of the growing number of studies that have been carried out. Indeed, there are  
 103 other reviews which attempt to do this [13, 17, 21-23]. Rather, we focus on key areas of interest,  
 104 encompassing an overview of approaches with practical applications and priority needs. The focus  
 105 here will be (i) to cover the different areas of interest to managers, (ii) to provide a brief overview of  
 106 eDNA-based methods and strategies and (iii) to outline their state of development, practical uses, and  
 107 development requirements, together with their limitations and factors which need to be addressed  
 108 when integrating these tools into the management of marine resources.



111 **Fig. 1.** Different methods for sampling marine ecosystems associated with their DNA source, type  
 112 of sample obtained and target organisms. Target organisms are shown based on the source of the  
 113 DNA collected.

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 115 **2. Environmental DNA in a fisheries context**

116 The marine environment harbours a huge diversity of species [24], ranging from large and  
117 charismatic whales to tiny worms and unicellular plankton (Fig. 1). Compared to the sampling of eDNA  
118 in freshwater it also poses its own set of, often difficult to address, issues when trying to obtain  
119 unbiased samples, especially in relation to factors such, tides, currents, great depths and rapid  
120 movements of individuals in three dimensions. Thus, depending on the habitat and taxa of interest,  
121 various sampling methods are needed to collect the full range of target species present at a given site  
122 so that, when possible, visual identification and quantification of the species is done to study, monitor,  
123 and provide information of relevance to the management of marine communities (Fig. 1).

124 Identification and characterization of these samples can be accelerated using genetic  
125 techniques. These will differ depending on the source of the DNA obtained. In the first case,  
126 community DNA can be collected. This refers to the collection of whole communities of organisms in  
127 the sample from which DNA is extracted from the cells of the sampled individuals. Such analysis results  
128 in highly comparable results for monitoring and impact assessment, compared to traditional  
129 morphological analyses [25, 26] and at a fraction of the time and cost [25]. In the second case,  
130 organisms are not directly sampled, rather extraorganismal DNA in the environment (eDNA) is  
131 collected and used to infer a species presence. The use of eDNA in this way may even further simplify  
132 sampling and increase throughput, decreasing the costs and allowing for large scale surveys of marine  
133 ecosystems.

134 Traces of DNA in the water column and in the sediment can be used to identify species and  
135 characterize communities [e.g. 27], to investigate their distribution [e.g. 28], and to determine their  
136 abundance [e.g. 29]. Both community DNA and eDNA data are affected by technical (e.g. laboratory  
137 assay choices, incomplete reference databases) and biological (e.g. size of the organisms) biases,  
138 which should be taken into account when interpreting the data for fisheries management and  
139 ecosystem monitoring [30]. While the distribution of the entire organisms collected during community  
140 DNA surveys is, of course, affected by environmental parameters, extracellular eDNA is especially  
141 sensitive to such factors. eDNA data is thus influenced by environmental factors such as water



142 temperature, organic matter, pH, UV radiation, and water currents, and by the type and amount of  
143 material used during sampling [17]. Further, as eDNA is used as a proxy for species presence, any  
144 biases in the transport and persistence of eDNA can result in its distribution being significantly  
145 different from that of the actual organisms. Careful evaluation of these biases is needed for the correct  
146 interpretation of eDNA results in the framework of fisheries management and conservation.

### 148 **3. From water to results - the eDNA workflow and approaches**

149 Identifying the presence of a particular species or characterizing the entire community from  
150 eDNA samples requires a series of steps that often need to be adjusted to each case study and fully  
151 understood in order to derive sound conclusions from the data obtained [30]. Sampling eDNA in the  
152 marine environment is possible through water or sediment [31]. It is however usually done by  
153 collecting water that is subsequently passed through variable pore size filters, generally < 1 µm pore  
154 size. It is also often common practice to add a prefiltering step (e.g. with a 3 µm prefilter) to avoid  
155 clogging the filtering process with large pieces of tissue or small animals such as zooplankton [32].  
156 Water samples from the marine environment can be collected using procedures that span from the  
157 simple act of using a bucket to collect surface samples to a more sophisticated procedure involving  
158 the use of Niskin bottles [33] or rosette samplers [34] to capture samples at greater depths. In all  
159 cases, strict procedures to avoid cross-contamination between samples are needed along with proper  
160 preservation and storage for filters containing eDNA prior to laboratory analysis. While applications  
161 are diverse, approaches using eDNA can be categorised into three groups based on their main  
162 objectives: 1) *Targeted Species Detection*, to detect the presence or absence of a single or a limited  
163 number of defined targeted species at a location; 2) *Community Characterisation*, to produce an  
164 inventory of the biodiversity of an ecosystem; and 3) *Species Abundance Estimation*, to inform on  
165 absolute and/or relative abundance of species at the sampling location. An overview of the three  
166 groups is presented below, detailing their objectives, strengths and limitations. Selected examples of  
167 each technique are also outlined in Tables 1-3 to show typical situations where they have been utilised.

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169 3.1. Targeted species detection

170 Perhaps the most developed and utilised eDNA application is the detection of individual species  
 171 and/or small groups of targeted species of interest in an ecosystem. Targeted species detection from  
 172 eDNA involves the development of genetic probes designed to match explicitly the target species DNA,  
 173 and distinguish the target from other species potentially present in a sample using classical genomic  
 174 Sanger sequencing [13, 35, 36] and/or quantitative real time PCR (qPCR) [37]. Marker amplification is  
 175 achieved by the use of DNA probes, which allow the genetic code of specific sections of the genome  
 176 to be examined, and resulting unique species-specific genetic sequences. qPCR is based on detection  
 177 and quantification of a fluorescent light signal produced by binding of a dye-labelled species-specific  
 178 probe, during amplification, to the target species DNA sequence present in a sample [38]. Detection  
 179 of small groups of species using qPCR can be achieved by combining (multiplexing) probes for these  
 180 species, labelled with different fluorescent dyes, in a single reaction.

181

182 **Table 1**

183 Selected applications of targeted species detection using marine eDNA.

184

Application	Example study outline	Example
Detection and mapping of the spread of invasive or non-native species	Invasive slipper shell on the European Atlantic coast	[39]
Identification and monitoring of rare/endangered species	White sharks in the open ocean	[34]
Detection of cryptic species	Cryptic seahorse species off western Australia	[40]
Biosecurity during import/export	Ornamental fish imports	[41]
Investigating spawning activity	Spawning ecology of the Japanese eel	[42]
Monitoring of hard to access environments	Deep-sea octocorals using remote submersibles	[43]

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186 Applications are varied and are detailed with examples in Table 1. It can be observed from these  
187 examples that targeted species detection has shown its usefulness across many and varied situations  
188 of fishery management and ecosystem monitoring. Marine monitoring using traditional methods such  
189 as individual capture (with e.g. trawls, nets and traps) and visual surveys are time consuming, costly  
190 to carry out and in some cases simply impossible. Investigations using eDNA have shown that in  
191 numerous situations the approaches have the potential to add to the available information to inform  
192 a variety of management questions. Adding value to traditional programmes is, perhaps, the most  
193 cost-effective way to integrate eDNA screening into routine management and monitoring  
194 programmes (see below). However, in some specific situations the use of eDNA has the potential to  
195 replace traditional monitoring. For this to occur a number of technical and validation steps are  
196 required such as comparisons between eDNA and visual survey data in context, controls for type I  
197 (false-positive) and type II (false negative) errors, validation of experimental results in the laboratory,  
198 scaling up versus one-off sample collection, temporal and spatial replicates (see below). If such steps  
199 are successful, targeted species detection using eDNA has shown that it can fulfil the requirements of  
200 fishery and ecosystem monitoring programmes and can be used as an alternative approach to answer  
201 relevant questions for managers.

**Box 1. Case study – Targeted species detection – eDNA and ecology of commercially important food species [42]**

The catadromous Japanese eel *Anguilla japonica* is an important food fish in East Asia, where after spawning at sea and migrating to freshwater it is raised in aquaculture ponds. Intensive research including sampling with large plankton and trawl nets, genetic species identification of eggs and newly hatched larvae, and direct observations using deep-tow camera systems has led to the discovery of the eel's spawning area. Such approaches have provided useful information on the spawning area of Japanese eels. However, their precise spawning sites and ecology still remain largely unknown, in part due to the significant depths and vast scale of the possible survey areas and the need to narrow down the search areas.

213 In order to address these issues, species-specific genetic probes were developed and tested in  
214 the laboratory by filtering and extracting eDNA from tank water containing eels. This showed that the  
215 probes could identify the Japanese eel from a minute amount of eDNA. Samples were collected at  
216 varying depths during an ocean survey on the southern West Mariana Ridge in the general spawning  
217 area of the eel. eDNA positive signals were detected for *A. japonica* from 3 of the 108 samples.

218 This first attempt to detect Japanese eel eDNA suggests the approach has the potential to  
219 provide information in near real-time about the spawning aggregations in a deep-water environment  
220 which is very challenging to survey using traditional techniques.

### 222 3.2. Community characterisation

223 Community characterisation, often referred to as community metabarcoding, is a technique  
224 used to characterise either the species composition or a selected subset of species, whose eDNA is  
225 represented in a water sample [44, 45]. Using this approach, a region of DNA conserved within a  
226 species and diverse across a wide range of taxa is specifically targeted and many targets are captured  
227 simultaneously in a single reaction. Amplified products are sequenced, revealing unique species-  
228 specific signatures (i.e. a barcode for that species) within a sample and sequences are compared to  
229 reference sequences within a database. As such, each unique sequence match between the sample  
230 and the reference database will identify DNA from a specific species in the sample [46]. Metabarcoding  
231 has been utilized in a variety of settings, showing a broad potential application for biodiversity  
232 monitoring (Table 2).

234 **Table 2**

235 Selected applications of community characterisation using marine eDNA.

236 Application	Example study outline	Example
Fish diversity	Fish community composition in a large (120,000 km <sup>2</sup> ) area of the NE Atlantic	[47]

1	Identification of new species in an area	Detection of a number of invasive, cryptic and observations of species for the first time in the North Sea	[48]
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4	Connection of life stages	Linking distributions of adult and immature stages of South African marine fish species	[49]
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8	Clarification of feeding behaviour	Characterisation of prey species of invasive lionfish through gut content analysis in the Mexican Caribbean	[50]
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12	Ecosystem food-web structure and dynamics	Characterisation of community structure of Japanese coastal waters	[51]
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15	The impact of aquaculture on benthic communities	Comparison of benthic Foraminifera communities at different distances from aquaculture sites	[52]
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19	Identification of non-indigenous species in ballast/harbour water	Detection of the transfer of North Sea molluscs across tropical waters in ballast water	[53]
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23	Monitoring of marine vertebrates	Distribution in space and water column of marine vertebrates in Monterey Bay	[54]
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26	Habitat preference	Fine-scale geographic and temporal mapping of marine fish populations in the Hudson River estuary	[55]
27			
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29			
30	Characterisation of non-indigenous species	Detection of introduced and newly observed resident marine species around southern Britain	[27]
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34	Biodiversity assessment- marine sanctuaries	Characterisation of pelagic and benthic eukaryotic biodiversity in the Florida Keys National Marine Sanctuary	[56]
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238 eDNA metabarcoding is well established in providing unique insights into the diversity and  
 239 functioning [57] of aquatic ecosystems. Such applications have allowed the characterisation of fish  
 240 communities in freshwater [e.g. 58] and marine [e.g. 59] environments, including pelagic [e.g. 60] and  
 241 benthic communities [e.g. 61]. Together with such an often-unique ability to characterise entire  
 242 communities, metabarcoding has also been used in a more applied way to answer specific questions  
 243 of interest to managers and policy makers. These include investigations of the impact of aquaculture  
 244 on local bottom communities, the transfer of non-indigenous and invasive species in ballast and  
 245 harbour water, and monitoring of marine vertebrates (Table 2). Where targeted species detection  
 246 using eDNA allows specific species to be examined, aquatic eDNA metabarcoding allows the cost-

247 effective characterisation of entire communities, and therefore it is especially useful in ecosystem  
248 monitoring scenarios.

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**Box 2. Case study – Community characterisation – fish biodiversity assessment using eDNA over large oceanic areas [47]**

252 Traditional methods of monitoring marine fish diversity rely on trawling surveys. These are  
253 costly, time-consuming and, especially in complex environments, may be biased in the species they  
254 capture with only a sub-set being targeted. Community characterisation using eDNA has the potential  
255 to address some of these shortcomings by, in theory, being able to identify all species in an area using  
256 the eDNA they shed into the environment.

257 In order to test this hypothesis, an eDNA based metabarcoding approach was used to  
258 characterise the species present across a 120,000 km<sup>2</sup> area of the Northeast Atlantic using eDNA  
259 filtered from water samples. Species specific genetic sequences were obtained from the eDNA which  
260 were identified through matches in reference databases. The results of this analysis were compared  
261 to traditional trawl surveys carried out simultaneously to the water sampling.

262 It was found that trawl and eDNA samples resulted in the same most abundant species  
263 (European anchovy, European pilchard, Atlantic mackerel, and blue whiting), but eDNA  
264 metabarcoding resulted in more detected bony fish and elasmobranch species (116) than trawling  
265 (16). The eDNA metabarcoding approach was thus seen to capture the biodiversity present in the area  
266 at least as good, and with some groups of species better, than traditional techniques. The findings  
267 support the integration of eDNA metabarcoding for broad-scale marine fish diversity monitoring in  
268 the context of Directives such as the Common Fisheries Policy or the Marine Strategy Framework  
269 Directive.

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272 3.3. Species Abundance Estimation

273 Together with the identification of both individual and ecosystem-based biodiversity, eDNA can  
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 2 274 be used to estimate either the relative abundance of multiple species using metabarcoding [62], or  
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 4 275 the absolute abundance of individual species using qPCR [63]. At its simplest, such approaches involve  
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 7 276 quantifying the amount of eDNA from a species represented in a sample and using that as a simple  
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 9 277 proxy for abundance [64]. Such information may be used to estimate numbers of individuals and/or  
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 11 278 biomass. The use of eDNA-based tools to quantify stocks of species of interest is of course of great  
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 14 279 interest to fishery managers and policy makers, as population or stock assessment is a central  
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 16 280 component of any management and/or conservation programme. Estimating absolute counts and/or  
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 19 281 biomass, relies on the establishment of a robust correlation between DNA concentration and living  
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 21 282 biomass whereas relative biomass estimates assume that the relative amounts of DNA measured in  
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 23 283 the sample are representative of the relative abundance of the different species in the ecosystem.  
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 26 284 While both approaches may seem to rely on fairly simple calculations and indeed are beginning to be  
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 28 285 used (Table 3), in practice, there are many factors which interact to make the relationships upon which  
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 31 286 the assumptions about the correlations are made very complex to disentangle and to obtain robust  
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 33 287 estimates.

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37  
 38 **Table 3**

39  
 40 290 Selected applications of abundance estimation using marine eDNA.

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Application	Example study outline	Example
Seasonal fish abundance	Seasonal relative fish species abundance in the Hudson River estuary	[55]
Marine vertebrate abundance	Vertebrate relative abundance in a kelp forest off the Monterey Peninsula	[65]
Monitoring pathogen abundance in aquaculture	Relative abundance of two parasite species on salmon farms	[66]
Monitoring deep water species	Relative abundances of Subarctic, deep water fish species from the continental slope off Southwest Greenland	[62]

Invasive species abundance	Temporal abundance of invasive Codium seaweed in the Bay of Biscay	[67]
Stock assessment	Biomass estimation of Atlantic cod in oceanic waters around the Faroe Islands	[29]

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Applications of using eDNA to assess abundance in the aquatic environment are at present most advanced in freshwater [62]. Abundance estimation using traditional methods such as gillnet data and trawling provides a relative index assumed to be directly proportional to density/absolute abundance [29, 64, 68]. Such traditional non-genetic methods are the most common to estimate fish abundance in lakes for fisheries management [69] and biodiversity characterisation [70], although they are often expensive, time consuming and destructive. Initial results from experimental aquaria and ponds show positive correlations between species abundance and eDNA concentration [71, 72]. However, even in controlled tank situations, it has been found that “...quantification of eDNA samples can be highly variable even when sampling from the same individual under controlled conditions” [72]. Approaches have now moved from the experimental set-up to the field. The abundance of individual targeted species has been characterised using eDNA in freshwater fish species including lake trout (*Salvelinus namaycush*) [64], common carp (*Cyprinus carpio*) [73] and Atlantic salmon (*Salmo salar*) [74]. Similarity between relative and absolute abundance has been reported in communities including both amphibians [75] and fish [55, 76], including commercially important species such as Atlantic cod (*Gadus morhua*) [29].

**Box 3. Case study – environmental DNA and quantitative assessment of commercial fish species [29]**

Traditionally, standardised trawl surveys are used as an effective monitoring tool for management of commercial fisheries, providing valuable estimates of quantity (biomass) and spatial distribution of fish stocks. Such surveys, however, are costly and have other associated biases and drawbacks such as gear and ground selectivity and negative impact on habitats.



314 In order to determine the utility of eDNA for assessing commercial stocks a quantitative eDNA  
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2 315 survey of Atlantic cod was compared to results from a standardised demersal trawl survey. Important  
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5 316 stock metrics such as regional cod biomass and Catch Per Unit Effort (CPUE) were determined using  
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7 317 traditional assessment analysis of trawl data. At 35 trawl stations water samples were also collected  
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9 318 4 m above the seafloor and eDNA analysed in the laboratory using cod-specific DNA probes.

11 319 There was an overall 80 % concordance between trawl and eDNA cod detection, with good  
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14 320 spatial conformity between the two approaches. Nearly 70 % of all discrepancies in the detection of  
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16 321 Atlantic cod were at the sampling stations where actual or predicted Atlantic cod catch rates were  
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19 322 very low ( $\leq 3$  fish  $h^{-1}$ ). Similarly, there were also significant positive correlations between the regional  
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21 323 integrals of cod biomass (kg) and eDNA quantities (copies) and between sampling effort-normalised  
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23 324 CPUE and eDNA concentrations.

26 325 This study shows that eDNA monitoring can provide valuable spatial and abundance  
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28 326 information which is comparable to traditional standardised trawl data but less costly and with less  
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31 327 impact on the environment. The findings reinforce the opportunities for the incorporation of  
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33 328 approaches utilising eDNA into stock biomass assessments of commercially important fish stocks.  
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38 330 In the marine environment, abundance estimates using eDNA, while inherently more difficult  
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40 331 than a relatively enclosed freshwater ecosystem, are starting to be examined (Table 3). Approaches  
41  
42 332 are developing rapidly and, while at present robust relationships between abundance quantification  
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45 333 using eDNA and more traditional methods are sometimes weak [62, 77, 78], in some cases the  
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47 334 approach seems to be comparable to that of other quantitative methods [29, 79]. The inherent  
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50 335 uncertainty in the robustness of biomass quantification when utilising eDNA approaches is due to both  
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52 336 the assumptions on which the technique rests and the impact of extraneous factors on such  
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54 337 assumptions. eDNA abundance quantification relies on the assumption that local population numbers  
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57 338 may be inferred by measuring the concentration of eDNA at a given locality and that this estimation  
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59 339 represents the quantitative relation between eDNA concentration and the underlying population size  
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340 [79, 80]. However, such a relationship may not be always true, or even present in most cases. The  
341 amount of eDNA at a location will vary depending on a number of biological, physical and  
342 environmental factors (see below). While these factors also have an impact on species detection, the  
343 impact of the fluctuations registered is higher if quantitative measurements are being attempted,  
344 rather than simple presence/absence results. Nevertheless, it may be possible to incorporate these  
345 impacts into modelling, to better predict how they can affect eDNA concentrations, therefore  
346 reducing the variance around such quantifications [79, 81-83]. However, due to the complexity of  
347 interacting factors, direct quantitative assessments remain highly challenging in marine ecosystems  
348 [17, 84].

349 Abundance estimates in the marine environment can thus be summarised to be very much in  
350 the developmental stage at the moment, notwithstanding some of the early applications being  
351 examined. Significant questions still have to be addressed to allow the amount of eDNA collected to  
352 be linked directly to either relative or absolute abundances. The three-dimensional nature of the  
353 environment, together with the many physical, chemical and environmental factors whose impacts  
354 have to be quantified means that the validity of abundance quantification using eDNA is still to be  
355 determined in most if not all situations. Significant work is, however, being undertaken around the  
356 world to determine if the method can be developed into a useful tool as, if so, it might in the future  
357 provide a very cost-effective approach. At present, however, the jury is still out if this will be possible.

358

#### 359 **4. Considerations**

360 Analysis of eDNA allows inferences to be made about organisms, without the need to see,  
361 observe or handle them. This is the major advantage offered by this approach, but also potentially a  
362 drawback. In order to make the most informed decisions and use eDNA approaches to their fullest,  
363 managers and policy makers should be aware of the issues to be considered when seeking to  
364 understand the results of eDNA surveys. Although eDNA based applications are relatively new,  
365 especially in the context of marine management, scientists have a good understanding of the

366 drawbacks of this method, hence have been able to define the actions needed in order to limit errors  
1  
2 367 and uncertainties [85-87].  
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4 368 An important consideration in any eDNA monitoring programme is the avoidance of  
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7 369 contamination [88]. DNA molecules from many sources are everywhere around us, and if they enter  
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9 370 eDNA samples they have the potential to produce false positives. The use of sterile equipment, gloves,  
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11 371 and a dedicated eDNA laboratory (with strict protocols, controls and necessary separations of  
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13 372 processes handling high and low DNA templates) are necessary measurements to be taken in order to  
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15 373 reduce contaminations and resulting false positives [86]. It is possible to control for contamination, by  
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17 374 taking multiple replicates (usually three) of the same samples, and by using negative controls (i.e.  
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19 375 sterilised distilled water samples not containing any actual material) at every stage of the process  
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21 376 (field and laboratory blanks for DNA extraction and amplification) [88]. Any DNA that results from  
22  
23 377 these blanks (and there is likely to be some), is then 'subtracted' from the results of the actual samples.  
24  
25 378 Thus, like in any other monitoring approach, standardization is crucial, especially when it comes to  
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27 379 techniques of collection, essential negative control sample inclusion [89] and laboratory analysis [90],  
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29 380 as well as the interpretation of results [91].  
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35 381 Another important consideration (which can be a significant drawback in certain situations) is  
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37 382 the availability of DNA reference sequences, or a reference database of taxonomically identified  
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39 383 species/groups [92]. Matching sequences obtained from actual eDNA samples against a reference  
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41 384 database is the final step in the workflow, one that will tell the user what species the sampled eDNA  
42  
43 385 belongs to. The reliability of such databases, together with the availability of high-quality reference  
44  
45 386 sequences of previously examined and taxonomically identified organisms is crucial for robust data  
46  
47 387 interpretation and to avoid false negatives and positives. There are a number of databases that can  
48  
49 388 be used, with the Barcode of Life Data System (iBOL) [93] being an important example. Yet, it is  
50  
51 389 advisable, when embarking on an eDNA project, to invest time assessing the reliability of the  
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53 390 databases for the geographic area and taxa investigated, and if required, build a project-specific  
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55 391 quality-controlled database.  
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392 Another pivotal consideration when interpreting results is that of eDNA transport. As  
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2 393 mentioned above, eDNA offers a snapshot of the species presence in a certain habitat in a given  
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4 394 timeframe. Environmental DNA sampled might indeed come from the organisms that live in the  
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7 395 sampled area at that time, but it might also originate from degrading tissue, eggs and sperm and,  
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9 396 depending on environmental conditions, it might have simply been transported from elsewhere with  
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11 397 the currents or tides. Many researchers are now concentrating their efforts into understanding how  
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14 398 long these molecules can persist in the environment and remain detectable [reviewed in 17].  
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## 20 400 **5. Integration into existing management and monitoring programmes**

23 401 The development of new approaches to gather information of relevance to fisheries and  
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25 402 ecosystem monitoring through the use of eDNA sampling methods, and the associated novel insights  
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28 403 such approaches generate, has the potential to revolutionise the information available to managers.  
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30 404 However, together with the requirement for the new methods to be able to provide robust results,  
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32 405 there is also a need to investigate the practicalities and cost-benefit of incorporating the new  
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35 406 techniques into standardised monitoring surveys [94, 95]. In some situations, for example, the  
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37 407 requirement for targeted detection of specific species, it may be necessary to develop novel surveying  
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39 408 programmes. However, by far the most preferred situation would be if the added value could be  
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42 409 embedded into existing survey programmes, through the addition of the collection of eDNA samples,  
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44 410 potentially requiring relatively little extra cost/effort on top of that already being invested. This is  
45  
46 411 especially relevant as ship-based survey costs increase while genetic screening costs are decreasing.  
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48 412 Trawl surveys may be able to be supplemented by simultaneous eDNA collection from water samples,  
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51 413 and benthic sediment monitoring by eDNA collection from grab samples. Indeed, in many if not most,  
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53 414 often costly, traditional fishery and ecosystem monitoring surveys there would seem to be an ideal  
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55 415 opportunity to collect such samples and add value in this way. It seems, therefore, that the design of  
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58 416 future surveys, together with that of existing programmes, should be evaluated in the light of the  
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2 417 developments in eDNA approaches outlined above and the added value that the integration of these  
3 418 approaches could bring.

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7 420 **6. Conclusion**

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9 421 Rapid developments in the field of eDNA analysis have provided a range of new tools for  
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11 422 research scientists, and fishery and ecosystem managers. With such developments, it is not  
12  
13 423 straightforward for the manager to disentangle which tools can provide robust evidence to  
14  
15 424 incorporate into policy development discussions, and which are still in the developmental phase. In  
16  
17 425 tandem, reports about such advances in the mainstream media drive stakeholders to question  
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19 426 managers about the utility of the toolkits, including specific questions that might be difficult to answer  
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21 427 for a non-specialist. Here, we have attempted to provide a topic-based overview which goes some  
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23 428 way to address this problem, and thus can be of use to inform managers of the strengths and  
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25 429 weaknesses of the various approaches currently available.

26  
27 430 Environmental DNA-based tools have, for a number of years now, been providing reliable  
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29 431 evidence in areas such as single species detection, and the characterisation of ecosystem biodiversity.  
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31 432 As such, they represent a robust, cost-effective, and in an increasing number of cases a more sensible  
32  
33 433 option for managers and monitors for incorporation into their standard scientific toolkits. While  
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35 434 significant advances have been, and continue to be, made in the use of eDNA to quantify both relative  
36  
37 435 and absolute abundance, such analyses are less well developed and still suffer from uncertainties  
38  
39 436 associated with various environmental, biological and methodological challenges of these techniques  
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41 437 [17]. As these influences are studied and their impacts better understood such uncertainties will be  
42  
43 438 reduced. However, at present their application is likely to be more limited.

44  
45 439 Every scientific monitoring method has uncertainties and the field of eDNA research is no  
46  
47 440 exception. However, in many cases such uncertainty is well understood and as such, and considering  
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49 441 the potential significant benefits and potential cost-savings of the new tools available, managers and  
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51 442 monitors should consider the integration of these approaches in their management planning

443 discussions along with the more traditional techniques. The different approaches can work together  
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2 444 to provide complementary information. In the end they will allow enhanced scientific understanding,  
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4 445 resulting in improved science-based policy development in view of ecosystem-based management.  
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## 9 447 **7. Acknowledgments**

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## 16 450 **References**

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