

The wild cost of invasive feral animals worldwide

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Highlights:

1. Invasive feral animals cost \$141.95 billion globally, mainly impacting the agriculture sector.

2. Oceania, North America, and Europe recorded the highest economic burden

3. Islands face a higher economic cost (\$83.40 billion) than mainlands, due to invasive feral livestock.

4. Damage costs outweighed management and mixed costs, totalling \$124.94 billion.

32 5. Invasive feral species cost twice as much as their wild invasive species counterparts.

33

34

35 **Abstract**

36 Invasive non-native species are a growing burden to economies worldwide. While domesticated
37 animals (i.e., livestock, beasts of burden or pets) have enabled our ways of life and provide
38 sustenance for countless individuals, they may cause substantial impacts when they escape or are
39 released (i.e., become feral) and then become invasive with impacts. We used the *InvaCost* to
40 evaluate monetary impacts from species in the *Domestic Animal Diversity Information System*
41 database. We found a total cost of \$141.95 billion from only 18 invasive feral species. Invasive
42 feral livestock incurred the highest costs at \$90.03 billion, with pets contributing \$50.93 billion
43 and beasts of burden having much lower costs at \$0.98 billion. Agriculture was the most affected
44 sector at \$80.79 billion, followed by the Environment (\$43.44 billion), and Authorities-
45 Stakeholders sectors (\$5.52 billion). Damage costs comprised the majority (\$124.94 billion), with
46 management and mixed damage-management costs making up the rest (\$9.62 and \$7.38 billion,
47 respectively). These economic impacts were observed globally, where Oceania, North America
48 and Europe were the most impacted regions. Islands recorded a higher economic burden than
49 continental areas, with livestock species dominating costs more on islands than mainlands
50 compared to other feral species. The costs of invasive feral animals were on average twice higher
51 than those of wild species. The management of invasive feral populations requires higher
52 investment, updated regulations, and comprehensive risk assessments. These are especially
53 complex when considering the potential conflicts arising from interventions with species that have
54 close ties to humans. Effective messaging to raise public awareness of the impacts of feral
55 populations and appropriate legislation to prevent or control such invasive feral populations will
56 substantially contribute to minimizing their socioeconomic and environmental impacts.

57

58

59 **Keywords:** biological invasions; agriculture; non-native species; economic impacts; conservation;

60 *InvaCost*

61 **Introduction**

62 For millennia, humans have domesticated animals which have improved well-being and advanced
63 cultural and socio-economic development (Diamond 2002; Zeder, 2012). Domestication refers to
64 the adaptation of animals to circumstances defined by humans and the establishment of a symbiotic
65 relationship, although a clear and comprehensive definition is still debated (Kohane & Parsons,
66 1988; Zeder, 2012; Purugganan, 2022). The earliest known domesticated animals were dogs
67 *Canis lupus familiaris* more than 13,000 years ago (Sablin & Khlopachev, 2002), likely first tamed
68 for hunting and protection, and goats *Capra hircus* to produce milk, meat, and wool (Ahmad et
69 al., 2020). During the period of maritime exploration, sailors introduced domesticated animals
70 such as goats and pigs to islands as a future food resource (Cheke, 2010). Cats were also introduced
71 inadvertently, as hitchhikers on ships and often settling on these islands (Whinam, 2005). The
72 global reliance on such species for food and economic security has since led humans to translocate
73 increasing numbers of domestic species favoured in livestock farming around the world (Crosby,
74 1986; Bradford, 1999; Rostagno, 2009). Currently, it is estimated that humans, together with their
75 livestock, comprise approximately 96% of terrestrial mammalian biomass on Earth, with profound
76 implications for the environment (Bar-On et al., 2018). Furthermore, poultry biomass (principally
77 the chicken *Gallus gallus*) is estimated to be three times higher than that of wild, i.e., non-
78 domesticated birds (Bar-On et al., 2018). This pattern of reliance is expected to continue, with a
79 sharp increase in the *per capita* consumption of animal-based food items by 2100, predominantly
80 in low-income countries (Bradford, 1999; Rohr et al., 2019).

81

82 Non-native species are commonly defined as those which are translocated by human
83 activities outside of their native range and where they have not naturally evolved (McNeill, 2003;
84 Soto et al., 2023a). These non-native species then—in some cases—establish wild populations,
85 spread, and cause negative ecological, economic, and/or social impacts (Blackburn et al., 2011).
86 However, this non-native definition can be challenging to apply to animals with a long history of
87 domestication (see Gurevitch & Padilla, 2004), owing to hybridisation events and adaptations that
88 have produced distinct lineages (McHugo et al., 2019). In this context, domestic animals present
89 a unique case: despite being domesticated and often introduced to new environments by humans,
90 they are generally not perceived as 'foreign'. Furthermore, some domestic species can be already
91 integrated into new ecosystems and thus considered as part of the natural environment (Gurevitch

92 & Padilla, 2004). This complicates their classification as either native or non-native by traditional
93 definitions (Gurevitch & Padilla, 2004; Moutou & Pastoret, 2010). In this study we refer to
94 populations of non-native domesticated animals as invasive feral species.

95
96 Despite the economic, social, and cultural benefits of domesticated species to their owners
97 or public when confined to their human-constructed ecological niches (e.g., in farmland, homes,
98 or aquaculture facilities; Purugganan, 2022), released or escaped domesticated animals that
99 become feral can become a growing threat to ecosystems, biodiversity and global economies
100 (Genovesi et al., 2012; Russell & Blackburn, 2017; Marra, 2019, but see Foley et al., 2005). For
101 example, dogs and cats pose a substantial threat to biodiversity as both pets and feral animals
102 (Doherty et al., 2017; Loss et al., 2022), prompting numerous debates around legislation and the
103 implementation of management efforts (Riley, 2019; Trouwborst et al., 2020; Oedin et al., 2021).
104 These ecological and socioeconomic impacts are particularly notable in insular habitats (Whittaker
105 & Fernández-Palacios, 2007; Bellard et al., 2017; Bodey et al., 2022), with cats in particular
106 driving population declines and extinctions of endemic vertebrates on more than 100 islands
107 worldwide (Medina et al., 2011, 2014). While the impacts of cats—particularly colonies of stray
108 cats—are massive and particularly severe on islands, obtaining reliable data on the ecological and
109 economic impacts in mainlands is more challenging due to the complexities involved in monitoring
110 and quantifying their effects in these environments (Trouwborst et al., 2020; Carrete et al., 2022).
111 Domesticated animals can also be important vectors of pathogens, including *Salmonella*,
112 *Toxoplasma*, and the influenza A virus and rabies, causing wildlife and human diseases, or even
113 death (Pauwels & Pantchev, 2018; Lycett et al., 2019; Johnson & Johnson, 2021). Furthermore,
114 domesticated species can cause significant damage to the agriculture sector (e.g., reduced crop
115 yields or decreased productivity), resulting in a massive economic burden (Smith et al., 2007;
116 McKee et al., 2020). These impacts may even surpass those from non-domestic species, which
117 may be assumed to have a higher potential for harming local communities. This assumption is
118 based on the idea that domestication typically involves selecting for favorable specific biological
119 traits for human use and thus loss of certain behavior patterns, whereas non-domestic species lack
120 this selective process (Price, 1984; Wright, 2015). Additionally, the extent to which a species' gene
121 pool has been modified during domestication could also influence its ability to thrive in natural
122 environments and thus their capacity to cause an impact (Price, 1984).

123

124 Despite the substantial economic impacts domesticated animals can have when they spread
125 beyond their intended anthropogenic environments (e.g., farms, homes), research into these
126 impacts has been limited to specific species or case studies, and no attempts have been made to
127 comprehensively quantify the monetary costs they cause (but see Legge et al., 2020). This has
128 resulted in estimates scattered across individual reports and studies that each have a narrow focus,
129 inhibiting the search for broadscale drivers. This lack of research obscures the full range of
130 monetary impacts that invasive feral species can have, hindering efforts to effectively manage and
131 mitigate their potential impacts. In particular, the inclusion of economic costs can raise societal
132 awareness about the risks these invasions pose (Diagne et al., 2020; Ahmed et al., 2023), and help
133 to develop more responsible management practices (Cuthbert et al., 2022a).

134

135 Here, we used the *InvaCost* database — the most comprehensive and robust database on
136 the economic costs of invasive alien species worldwide (Diagne et al., 2020, 2021) — to provide
137 the first analysis of the economic costs of invasive feral animal species included in the *Domestic*
138 *Animal Diversity Information System* (DAD-IS) database to date. The *InvaCost* database is
139 instrumental in identifying gaps in current knowledge and guiding future research priorities in the
140 field of invasive species management, enabling more effective strategies to mitigate the economic
141 burden caused by these species. Given the long history of domestication, and consequent
142 translocation, of a limited number of species, we expect: *i*) massive and widespread costs
143 worldwide, being most prominent for a few high-profile taxa; *ii*) that most costs will be incurred
144 through damages in industrial sectors such as agriculture while management expenditures will be
145 scarce; *iii*) that reported costs will be biased towards particular regions due to differences in
146 research effort, with particularly substantial discrepancies on islands; and *iv*), that the average cost
147 associated with non-domestic 'wild' species (i.e., invasive without any domestication history) to
148 be several orders of magnitude greater than invasive feral species due to a reporting bias in light
149 of human reliance on domesticated species.

150

151 **Methods**

152 To quantify the economic costs of domesticated species, we used the *Domestic Animal Diversity*
153 *Information System* (DAD-IS) database (last accessed on 20 January 2023) developed by the Food

154 and Agriculture Organization of the United Nations to identify domesticated animal species
155 (www.fao.org/dad-is). The main aim of DAD-IS is to support the conservation and sustainable use
156 of domesticated taxa due to their importance for human populations and cultures. The DAD-IS
157 database contains information on diverse breeds of domesticated animals worldwide (FAO, 2023).
158 The domesticated species identified in DAD-IS were classified into three categories:

159

160 I. *Livestock*: Also known as “farm animals”, are domesticated species used to produce a wide
161 variety of products for consumption, such as meat, milk, fur, or eggs. Some notable species
162 in this group are cattle *Bos taurus*, sheep *Ovis aries*, goats *C. hircus* and chickens *G. gallus*;

163

164 II. *Pets*: Species kept by humans for various reasons, such as companionship, protection,
165 entertainment or to provide emotional support. Some notable species in this group are dogs
166 *C. lupus* or cats *Felis catus*. We did not include animals which are kept by humans but
167 which are arguably not domesticated (e.g., fish, crayfish, rodents);

168

169 III. *Beast of burden*: Also known as “working animals”, are domesticated species that are used
170 to perform physical tasks, such as carrying or transporting goods, materials, or people.
171 Some notable species in this group are donkeys *Equus asinus*, camels *Camelus bactrianus*
172 and llamas *Lama glama*.

173

174 To synthesise the economic costs of invasive feral species worldwide, we used the latest
175 version of the *InvaCost* database (version 4.1) using the *getInvaCostVersion* function of the
176 *invacost* R package (Diagne et al., 2020; Leroy et al., 2020). *InvaCost* is a “living” database
177 that is regularly updated with new information and curated (e.g., removal of duplicate entries).
178 *InvaCost* entries were obtained using specific search criteria (i.e., systematic review) to collate the
179 costs associated with biological invasions (Diagne et al., 2020). This was enhanced by adding costs
180 from sources in more than 20 non-English languages (Angulo et al., 2021; Kourantidou et al.,
181 2023). This version contains 13,553 cost entries worldwide extracted from primary sources
182 (original research studies, reports, and other documents with direct information on the costs of
183 invasive species), and secondary sources (research articles, books, and other sources that
184 synthesize or review the available cost information).

185

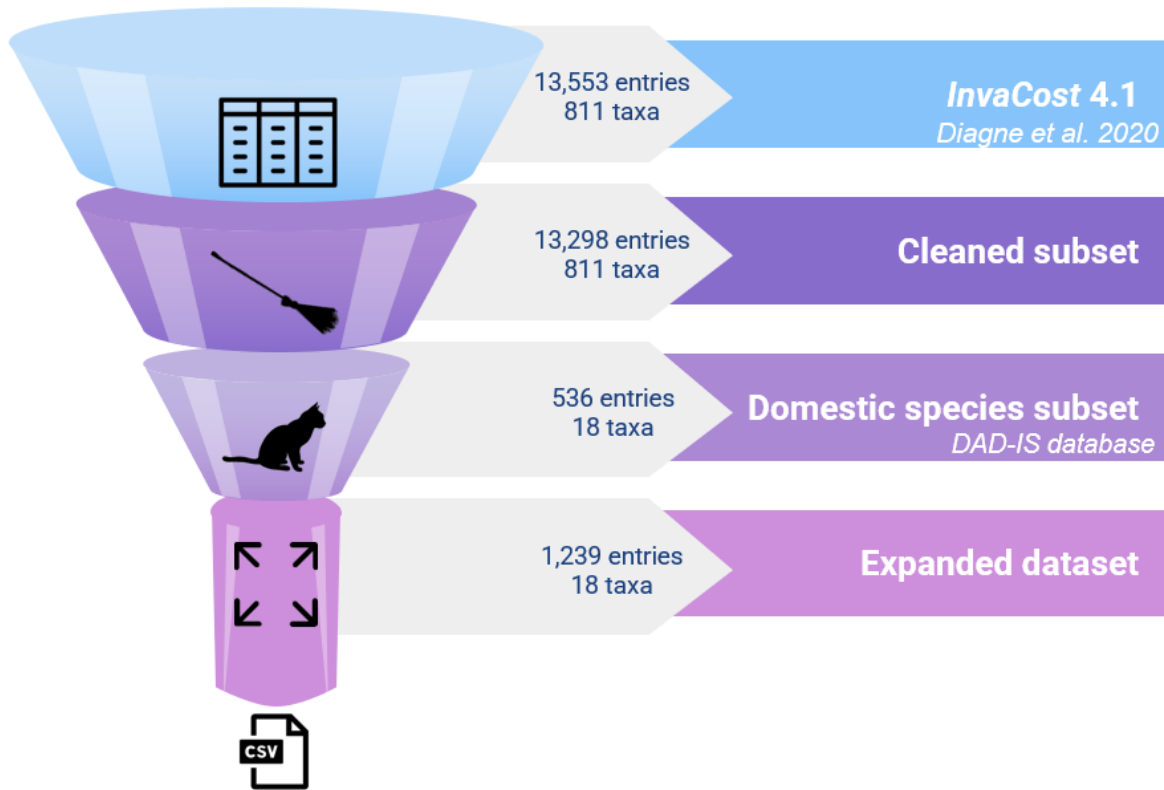
186 Each entry of *InvaCost* is standardised to a reference currency (US\$) and year (2017) to
187 control for inflation and allow for the direct comparison of costs over time and across currencies,
188 although since costs dating from before 1960 could not be standardised in this manner, they were
189 removed from our analyses. Each entry also includes a range of descriptors that provide additional
190 information about the cost, such as the specific species, impacted sector, region(s) where the cost
191 was incurred, and the methodology used to estimate the costs, among others (Diagne et al., 2020).

192

193 *Data processing*

194 Prior to analysing the data, we carried out a series of steps to clean and refine the database (Figure
195 1). We firstly removed entries that had incomplete temporal cost information, such as missing
196 starting or ending years (“Probable_starting_year_adjusted” and
197 “Probable_ending_year_adjusted”, respectively). We then identified species in DAD-IS that had
198 recorded costs in *InvaCost* by matching both databases based on the “Common_name” and
199 “Species” columns in *InvaCost* and “Common name” and “Scientific name” columns of DAD-IS.
200 Each species' name in DAD-IS was verified manually to ensure its accuracy to avoid any
201 misclassifications. Entries representing multiple species (e.g., *Cervus nippon/Dama*
202 *dama/Muntiacus reevesi*), which included one or more species not identified in DAD-IS, were
203 excluded from the analysis, due to the inability to disentangle individual species' impacts.

204



205
 206 **Figure 1.** Flowchart outlining the steps for selecting and processing data of domesticated species in *InvaCost* version
 207 4.1.

208
 209 The final dataset contained 536 entries from 18 taxa. To annualize the data, the total cost
 210 for each estimate and species was divided by the duration of the cost entry. This allowed the costs
 211 to be spread out over the years they occurred or were estimated to have occurred, ensuring their
 212 total value was not inflated. For example, a reported total cost of \$100,000 over five years would
 213 be transformed into five cost entries with a cost of \$20,000 per year. The duration of the cost was
 214 determined using the time between the “Probable_starting_year_adjusted” and the
 215 “Probable_ending_year_adjusted” columns, and this process was performed using the
 216 *expandYearlyCosts* function of the *invacost* R package (Leroy et al., 2020). This process
 217 resulted in a total of 1,239 annual cost entries (Figure 1). Each cost entry was analysed according
 218 to different cost descriptors as currently defined in the *InvaCost*: Method reliability,
 219 Implementation, Species, Impacted sector, Type of cost merged, Islands and Geographic region
 220 (see Supplementary Note 1).

222

223 *Feral vs. wild species comparison*

224 All invasive feral species considered fell into the taxonomic classes Mammalia and Aves.
225 Therefore, to test our hypothesis that invasive feral populations have lower economic costs than
226 those of non-domestic wild species, we compared the average economic costs between these two
227 categories across both originally domesticated (invasive feral species) and non-domesticated
228 invasive species (wild species). We obtained costs of relevant wild species by filtering the
229 tetrapods of *InvaCost* database (Bodey et al., unpublished data) considering only the relevant
230 classes, i.e., Mammalia and Aves (column Class), excluding costs spread across multiple species
231 where attribution to individual species was not possible, and finally excluding costs attributable to
232 invasive feral species. This resulted in a total of 56 wild species together with our dataset of 18
233 invasive feral species. Further, the average cost of non-domesticated and domesticated species was
234 calculated by dividing the total costs by the number of species identified in each category. In
235 addition, the costs in both non-domesticated wild and invasive feral species were split according
236 to their implementation and type of cost.

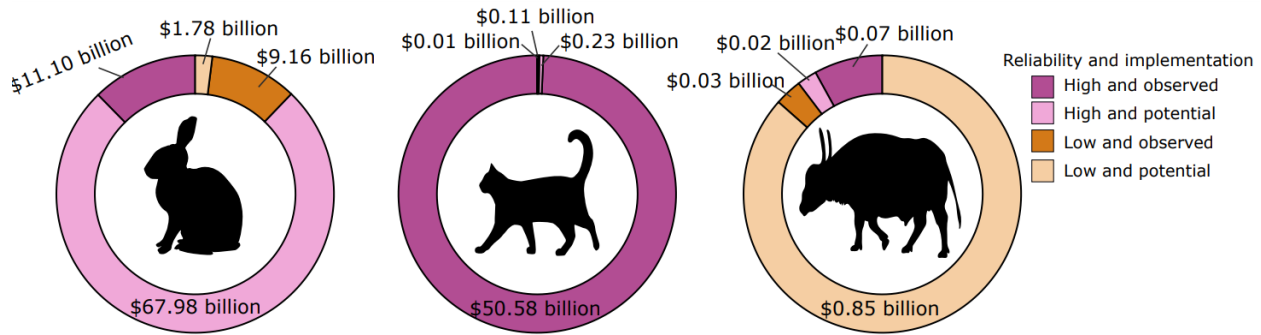
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238 **Results**

239 *Cost summary*

240 A total cost of \$141.95 billion (n = 1,239) arose from 18 invasive feral animals between 1960–
241 2022. Of this total, \$90.03 billion (n = 818) was attributed to 12 (out of 40 total) livestock species,
242 \$50.93 billion (n = 282) to the only two species categorized as pets by the DAD-IS (dogs and cats),
243 and lastly, \$0.98 billion (n = 139) to four (out of eight) beasts of burden (Table S1). For livestock
244 species, a large proportion of costs (\$79.08 billion, 87.83%) were from high reliability sources (n
245 = 677), whereas only \$10.95 billion (12.16%) were from low reliability sources (n = 141). Most
246 of the economic costs for pet species (\$50.83 billion, 99.80%) were from highly reliable sources
247 (n = 196), whereas only \$0.12 billion (0.20%) were from low reliability sources (n = 86). Of the
248 total costs (\$90.03 billion), most came from potential costs (\$69.76 billion, 77.49%, n = 271),
249 whereas only \$20.27 billion (22.51%, n = 547) came from observed costs. In terms of
250 implementation for pet costs, \$50.70 billion (n = 261) came from observed costs (99.5% of costs)
251 and \$0.25 billion (n = 21) from potential costs (0.50% of costs). The majority of costs for beasts
252 of burden were from low reliability sources (\$0.88 billion, 89.79%, n = 15), and \$0.10 billion

253 (10.21%) came from high reliability sources (n = 124). Most costs came from potential estimates
 254 (\$0.87 billion, 88.77%, n = 32), while a minor part was attributed to observed costs (\$0.11 billion,
 255 11.23 %, n = 107) (Figure 2).
 256

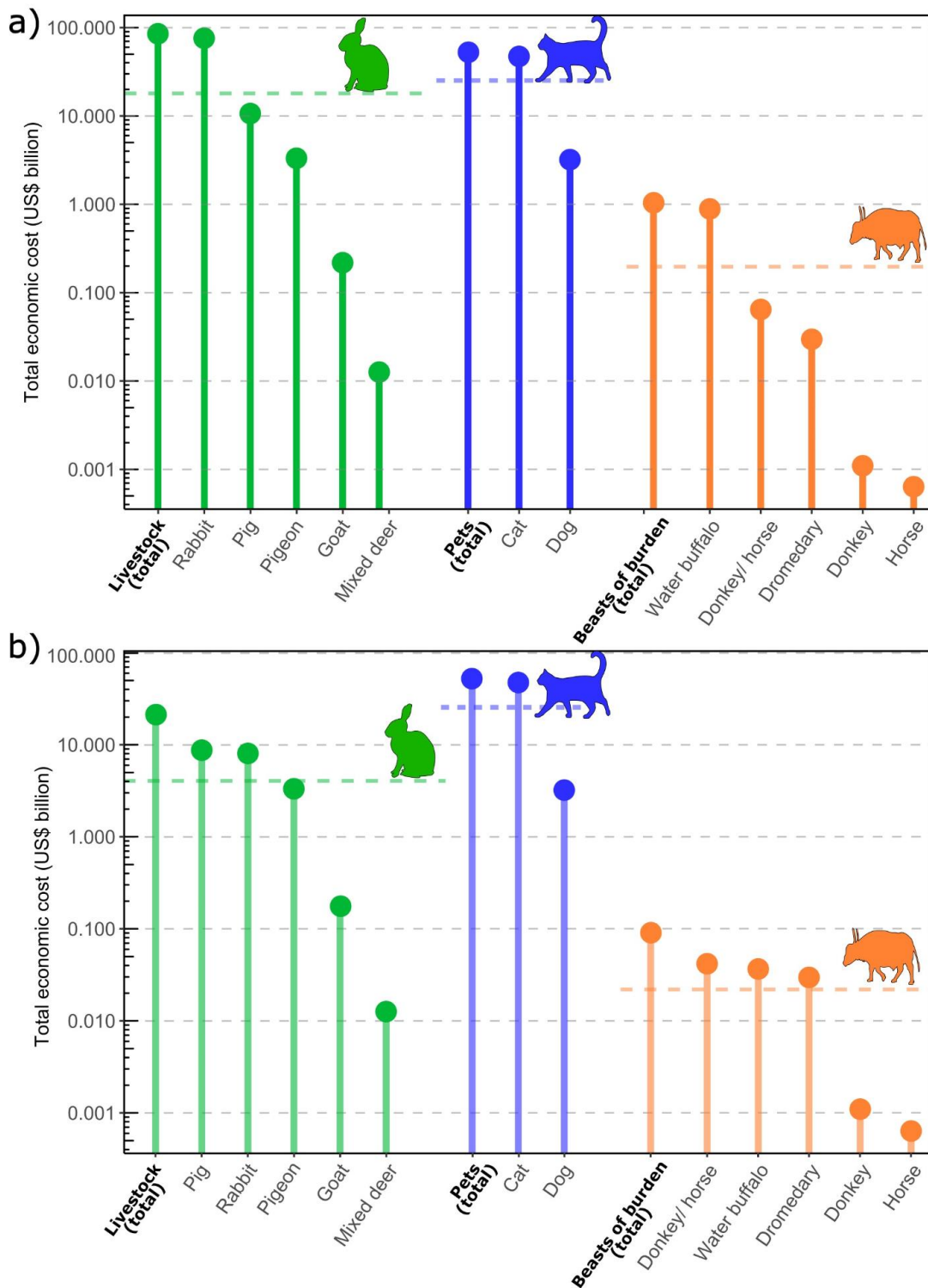


257
 258 **Figure 2.** The proportion (%) of total economic costs for each group of domesticated species (livestock, pets and
 259 beasts of burden) split into four categories according to method reliability (high or low) and implementation (observed
 260 or potential).
 261

262 *Economic costs among species*

263 The majority of costs (> 94.50%) was attributed to just three species: the European rabbit
 264 *Oryctolagus cuniculus*, the domestic cat, and feral pigs *Sus* spp. each exceeding \$10 billion. For
 265 livestock species, most of the costs (\$75.81 billion) were attributed to the European rabbit,
 266 followed by pigs (\$10.64 billion), pigeons *Columba livia* (\$3.32 billion), goats (\$0.21 billion), and
 267 mixed deer *Cervus nippon/Hydropotes inermis* (\$0.01 billion). For pets, the total was distributed
 268 between the domestic cat (\$47.70 billion) and dog (\$3.23 billion). For beasts of burden, costs were
 269 contributed from the water buffalo *Bubalus bubalis* (\$0.88 billion), mixed donkey *Equus asinus*
 270 and horse *E. caballus* (\$0.06 billion), dromedary *Camelus dromedarius* (\$0.03 billion), donkey
 271 (\$0.01 billion), and horse costs (\$0.0006 billion; Figure 3a).

272
 273 Considering only observed costs, pets superseded livestock as the costliest group overall,
 274 while the top five costliest invasive feral species remained the same, with pigs (\$8.73 billion)
 275 causing more costs than rabbits (\$8.01 billion). For beasts of burden the rank order changed, with
 276 water buffalo substantially reduced (\$0.03 billion) such that mixed costs of donkeys and horses
 277 (\$0.04 billion) were highest (Figure 3b).



278

279 **Figure 3.** Top-5 costliest species (US\$ billion) considering a) total economic costs (i.e., potential and observed costs

280 combined) and b) observed costs only for livestock (green), pets (blue) and beast of burden (orange) categories of

281 domesticated species following the respective group's total cost. The coloured dashed line refers to the average cost
 282 per species for each group of domestic species. Note that costs (y-axis) are on a log-scale.

283

284 *Impacted sectors and cost types*

285 Overall, we found that the sectors most impacted by invasive feral species were Agriculture
 286 (\$80.79 billion, n = 395), followed by the Environment (\$43.44 billion, n = 22) and Authorities-
 287 stakeholders (\$5.52 billion, n = 689). With respect to the type of cost, most costs (\$124.94 billion,
 288 n = 424) were categorized as damages, followed by management spending (\$9.63 billion, n = 797),
 289 and mixed damage-management costs (\$7.38 billion, n = 9; Table 1). Only \$0.005 billion (n = 9)
 290 was unspecified among types (Table 1).

291

292 **Table 1.** Monetary costs and numbers of database entries (rows) of groups of invasive feral species
 293 across impacted sectors and type of costs.

| Impacted sector | Cost in US\$ billion | Entries (n) | Domesticated group | Type of cost | Cost in US\$ billion | Entries (n) |
|---------------------------|----------------------|-------------|--------------------|--------------|----------------------|-------------|
| Environment | \$43.39 | 5 | Pets | Damage | \$45.04 | 40 |
| Health | \$5.30 | 9 | | Mixed | \$4.03 | 6 |
| Authorities-Stakeholders | \$1.85 | 233 | | Management | \$1.86 | 236 |
| Agriculture | \$0.24 | 33 | | | | |
| Public and social welfare | \$0.14 | 1 | | | | |
| Diverse | \$0.01 | 1 | | | | |
| Agriculture | \$79.62 | 326 | Livestock | Damage | \$79.84 | 344 |
| Diverse | \$4.58 | 50 | | Management | \$6.83 | 462 |
| Authorities-Stakeholders | \$3.64 | 410 | | Mixed | \$3.35 | 3 |
| Forestry | \$0.10 | 3 | | Unspecified | \$0.005 | 9 |
| Public and social welfare | \$0.07 | 4 | | | | |

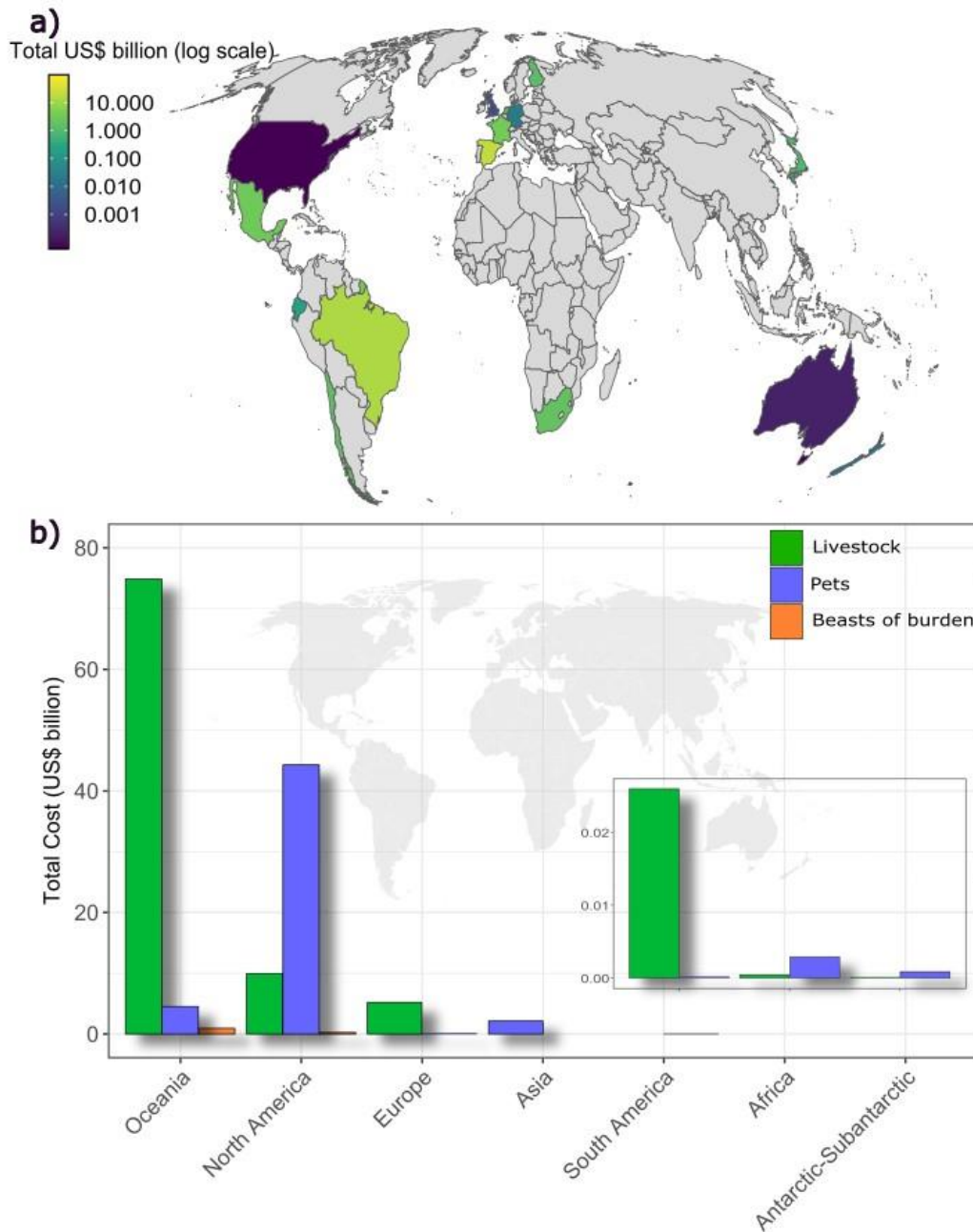
| | | | | | | |
|---------------------------|--------|----|-------------------------|------------|--------|----|
| Environment | \$0.05 | 17 | Beasts of burden | | | |
| Agriculture | \$0.93 | 36 | | Management | \$0.92 | 99 |
| Authorities-Stakeholders | \$0.03 | 46 | | Damage | \$0.05 | 40 |
| Public and social welfare | \$0.01 | 33 | | | | |
| Diverse | \$4.58 | 50 | | | | |
| Unspecified | \$1.95 | 8 | | | | |

294

295

296 *Geographic regions and islands*

297 Invasive feral species have global economic impacts, with costs recorded on all continents
 298 including Antarctica through sub-Antarctic islands (\$0.96 million, n = 8). However, the highest
 299 economic burden came from Oceania (including Pacific Islands; \$80.27 billion, n = 792), with
 300 almost all costs attributable to Australia, followed by North America (\$54.28 billion, n = 187, with
 301 the United States of America being the costliest country with \$54.25 billion) and Europe (\$5.21
 302 billion, n = 76, with the United Kingdom being the costliest country with \$3.01 billion), with all
 303 other regions incurring substantially lower costs (< \$2.22 billion; Figure 4). There was a notable
 304 gap in reported costs of feral populations in countries within Africa (with the exception of South
 305 Africa), Eastern Europe, the Middle East and Southeast Asia. For all animal groups, North
 306 America and Oceania incurred the highest costs: livestock, Oceania (\$74.81 billion, n = 523),
 307 North America (\$9.93 billion, n = 150); pets, North America (\$44.27 billion, n = 31),
 308 Oceania(\$4.52 billion, n = 136); beasts of burden, Oceania (\$0.94 billion, n = 133), North America
 309 (\$0.04 billion, n = 6).



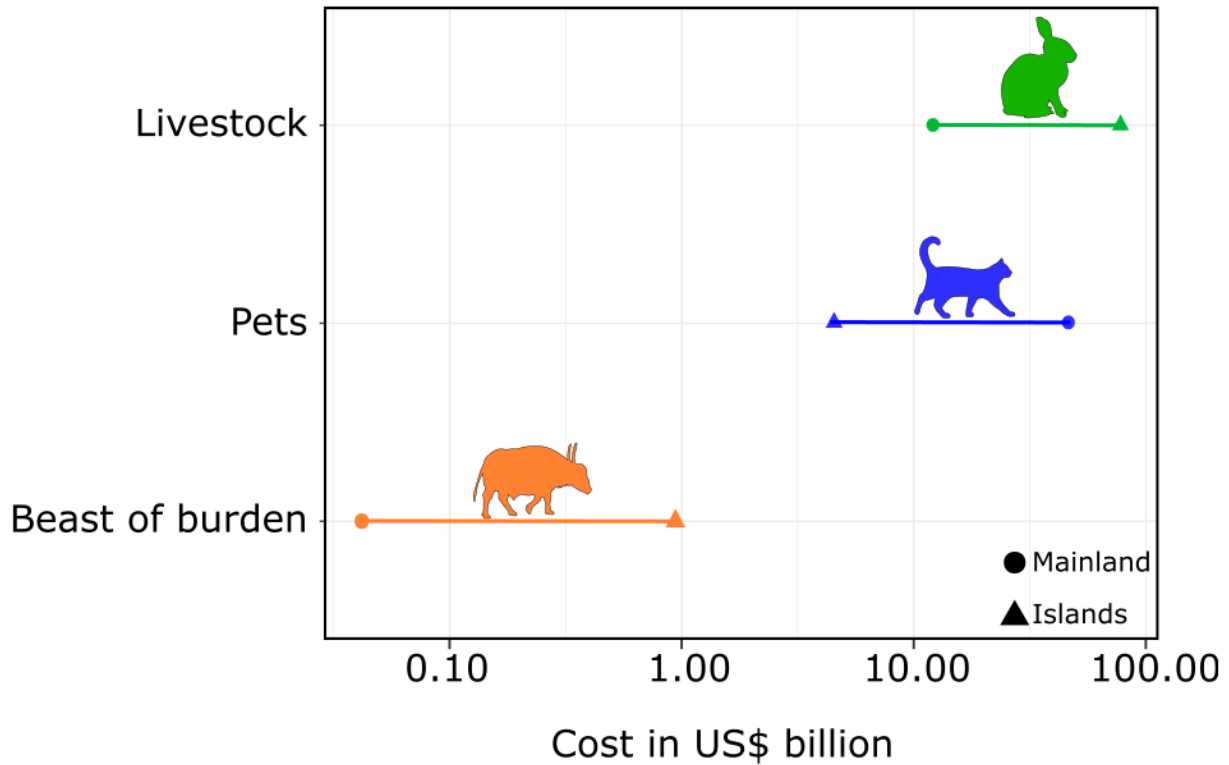
310

311 **Figure 4.** Geographical distribution of the economic costs of invasive feral species by country (a) and foreach group
312 of domesticated species (i.e., pets, livestock, and beast of burden) across geographic regions (b).

313

314 Islands recorded higher economic costs associated with domesticated species (\$83.40
315 billion, 58.73%, n = 1,087) than continental areas (\$58.55 billion 41.27%, n = 152). The greatest
316 proportion of costs on islands came from livestock species (\$77.92 billion; 93.43%), with costs
317 from pets totaling \$4.54 billion (5.44%). However, continental areas had contrasting ratios with

318 pets composing the greatest proportion at \$46.39 billion (78.74%), with livestock at \$12.11 billion
 319 (20.86%). While beasts of burden comprised only 1.05% of the total costs on islands at \$0.94
 320 billion, this figure comprised almost the entire recorded costs globally for this group (Figure 5).



321
 322 **Figure 5.** Total economic cost of each group of domesticated species (i.e., livestock, pets, and beasts of burden)
 323 between mainland and islands. Note that costs (x-axis) are on a log-scale.

324
 325 *Feral vs. wild species*

326 Compared to the \$141.95 billion total cost of invasive feral species across 18 species (\$7.88 billion
 327 per species), invasive wild species were associated with \$193.21 billion across 56 species (\$3.45
 328 billion per species). Interestingly, this means that invasive feral species were on average twice as
 329 costly as wild species, and this was not due to higher management costs. Indeed, the vast majority
 330 of the costs for invasive feral species and wild groups were due to damage (\$124.94 billion and
 331 \$166.13 billion, respectively), followed by management costs (\$9.62 billion and \$26.10 billion,
 332 respectively) and lastly, mixed costs (\$7.38 billion and \$0.92 billion, respectively). Albeit, the
 333 management investment for wild species was higher as a proportion of damage. In addition, a
 334 small portion of costs were classified as unspecified for invasive feral species and wild species
 335 (<\$0.1 billion). The percentage shares of observed and potential costs from invasive feral species

336 were similar, with \$71.06 billion observed (50.07%; \$3.94 billion per species), while higher shares
337 were reported for wild species at \$129.24 billion (66.89%, \$2.30 billion per species) in observed
338 costs.

339

340 **Discussion**

341 Although domestic species play an important role for human societies due to their contributions to
342 agriculture, companionship, and labor, as well as their cultural and economic value (Ahmad et al.,
343 2020), invasive feral populations of domesticated species have cost at least \$141.95 billion to the
344 global economy in the past 60 years. For livestock species, most of the costs were highly reliable
345 but potential (i.e., expected in the future), whereas for pets, most of the costs were both highly
346 reliable and observed. Beasts of burden, on the other hand, recorded the highest costs, which were
347 of low reliability and potential. Pets primarily impacted the environment, while livestock and
348 beasts of burden mostly affected the agriculture sector. Across all domestic groups, the most
349 common type of cost was damage. In terms of geographical distribution, the United States and
350 Australia stood out above other countries. Notably, only livestock and beasts of burden had a
351 greater impact on islands than on mainlands. Lastly, wild species caused twice the economic costs
352 as domestic species.

353

354 There was a notable disparity between the numbers of studies and associated costs for three
355 categories of domesticated species (i.e., livestock, pets and beasts of burden), with livestock
356 costing the most, followed by pets. This distinction may be attributed to the sheer numbers in
357 which livestock species are introduced and the expansive areas they occupy (Anderson, 2006).
358 Furthermore, their interactions with native species, competition for resources, and potential for
359 habitat degradation have a significant impact on recipient ecosystems, where they are often
360 considered “pests” and thus the target of eradication programs (Onuoha, 2008). While livestock as
361 a whole took the forefront in terms of costs, the majority of these expenses stemmed from the
362 European rabbit. Surprisingly, the economic toll from rabbits surpassed the combined costs from
363 all pets or beasts of burden.

364

365 *Method reliability and implementation*

366 Most of the economic costs and entries of invasive feral species were retrieved from highly reliable
367 sources, principally from governmental organizations. However, this was not the case for beasts
368 of burden, for which information was primarily retrieved from low reliability (i.e., grey literature)
369 sources. Together with the fact that the costs associated with beasts of burden were a very small
370 fraction of the overall costs associated with invasive feral species, this suggests that costs and
371 impacts have the potential to be substantially underestimated for these species as a result of both
372 the complexity of defining their native and invasive ranges, and their long-standing and intricate
373 history shared with humans (Crees & Turvey, 2015).

374

375 In terms of overall costs, an almost equal share was recorded as being directly observed
376 compared to potential incurred (i.e., extrapolated or predicted future costs). Potential costs were
377 higher than observed costs for livestock and beasts of burden, while the costs associated with pets
378 were mostly observed. Despite the massive costs for domesticated species, the large potential costs
379 being forecasted, particularly for livestock species, suggests that costs could increase further.
380 Potential costs, although uncertain, might not have fully manifested yet. This underscores the
381 importance of implementing improved monitoring and reporting mechanisms to track observed
382 costs in these species groups.

383

384 *Group differences in costs*

385 Although there were relatively high numbers of monetary costs associated with the two pets
386 included here, these costs are likely to still be substantially underestimated due to unquantified
387 costs associated with this group (e.g., snakes or fish) as well as substantial geographic and
388 taxonomic gaps in reported costs (Bush et al., 2014; Lockwood et al., 2019). Although most
389 domesticated individuals remain in captivity — and so far many species have not established
390 viable populations into the wild — there is a high risk of both escapes or releases of many species,
391 and thus potentially unquantified damages (Vall-Llosera & Cassey, 2017; Stringham & Lockwood,
392 2018). Indeed, such elevated levels of propagule pressure through the continuous introduction of
393 individuals increases the chances of successful establishment in the wild, which is further
394 facilitated by e.g., climate change and human alterations to the environment (Lockwood et al.,
395 2005, 2009; Kikillus et al. 2012).

396

397 Regarding livestock species, only 12 out of the 40 species included in DAD-IS (Table S1)
398 had recorded monetary impacts. Of them, the majority of costs were inferred by the European
399 rabbit. They are listed among “the 100 world’s worst” invasive non-native species due to their
400 massive impacts on ecosystems (Lowe et al., 2000). Such identification can lead to a greater
401 research focus, which may contribute to the high costs recorded for this species (Cuthbert et al.,
402 2022b). Most costs associated with rabbits were recorded as impacts on pastures and crops in
403 Australia (Gong et al., 2009). In some countries such as the United Kingdom, rabbits are
404 considered to be a ‘naturalised’ pest, rather than an invasive non-native species as a result of the
405 long timeframe since their original introduction, and such classifications can result in
406 underestimates of invasive species’ impacts (Cuthbert et al., 2021; Diagne et al., 2023). For beasts
407 of burden, four out of eight species in DAD-IS had recorded economic costs. The vast part of these
408 costs came from the water buffalo and were incurred in Australia (e.g., Ridpath & Waithm, 1988).
409 The relatively small pool of invasive beasts of burden can be due to the decreased likelihood of
410 detecting their impacts in regions with lower research effort, but also by the nuanced distinction
411 between native and invasive species. The association of domesticated species with beneficial
412 human activities (e.g., farming practices) can lead to shifted perceptions of “native” status for
413 domesticated species (Nance, 2015). For instance, species that have been part of agricultural
414 systems for centuries may begin to be perceived as "native" by the public in the context of human-
415 altered ecosystems, despite having impacts that we show to be akin to more conventional invasive
416 species when feral. While acknowledging that not all the species will cause a significant economic
417 burden, the scarcity of cost data across all groups and most of geographic regions highlights the
418 lack of quality data available for these groups. As a result, the estimates presented here are likely
419 conservative with regard to the full extent of their economic impact.

420

421 *Impacted sectors and cost types*

422 Introduction pathways play a defining role in the establishment and spread of non-native
423 populations (Turbelin et al., 2021). It is now recognized that when domesticated species
424 accidentally escape into the wild, they are able to survive, thrive and establish feral populations,
425 as for example, cats. This can lead to substantial impacts on various primary sectors such as
426 agriculture or forestry, alongside negative ecological consequences (Gong et al., 2009).
427 Paradoxically, agriculture, which often relies on domesticated species for its success, emerges as

428 the sector most negatively impacted by these feral populations. These impacts are principally
429 through crop damage, where overgrazing contributed to reductions in vegetation cover, soil
430 erosion, and loss of biodiversity (Filazzola et al., 2020). In addition, several management actions
431 can be hampered by pressure from different stakeholders, generating social conflicts around
432 conservation issues, control of invasive species that have become part of cultural practices in
433 invaded ranges, or those with socio-economic benefits (Massei et al., 2011; Crowley et al., 2017).

434

435 *Geographic distribution of economic costs*

436 Despite the widespread and massive economic costs of domesticated species, these were unevenly
437 spread globally, with costs particularly identified within Australia, the United States of America,
438 and the United Kingdom. This geographic pattern in cost reporting biases towards North America,
439 Oceania and Europe is a pervasive pattern reported on other invasive taxa (e.g., Haubrock et al.,
440 2021; Angulo et al., 2022), and likely reflects current and historic differences in research efforts
441 and economic activity (Nuñez et al., 2022). Countries characterized by growing economies and
442 substantial dependence on agricultural practices are anticipated to confront formidable challenges.
443 These countries must allocate resources towards domesticated species-centric production
444 endeavors to uphold food security. However, they face heightened vulnerability to the potential
445 collapse of their industries resulting from the impacts of invasive alien species (Turbelin et al.,
446 2023).

447

448 For livestock species, the largest share of economic costs originated from Oceania,
449 particularly from rabbits in Australia (Vere et al., 2004). The economic costs of livestock species
450 in other geographic regions were a share of ~17%, suggesting geographic knowledge gaps with
451 e.g., only 13 cost entries in Africa. These knowledge gaps are accentuated for beasts of burden,
452 with only two geographic regions having available cost data (North America and Oceania). The
453 reason for these knowledge gaps could be traced back to the history of domestication, which is
454 intrinsically linked to colonialism. Most of the domestication processes for these species began in
455 the Eurasian region, influencing the geographical distribution of economic costs associated with
456 them. Consequently, our understanding and quantification of the economic impacts are confined
457 to specific regions, neglecting the potentially considerable impacts in under-studied areas.

458

459 The impacts of domesticated species are particularly important on islands due to their
460 unique and vulnerable ecosystems. Among domesticated species, rabbits and cats are the most
461 impactful on islands — being also the driver of many animal extinctions (Medina et al., 2013). For
462 example, those impacts caused by cats on islands, where they have contributed to the extinction of
463 >60 native species (Meli et al., 2010; Doherty et al., 2016), are rarely translated into monetary
464 costs, or if translated, typically comprise relatively low management costs rather than a record of
465 damages to resources (Cuthbert et al., 2022a). Many other domestic species have caused drastic
466 damage to the islands invaded such as goats, pigs and sheep (Courchamp et al., 2003). These
467 species cause habitat degradation by reducing the vegetation cover, loss of native biodiversity
468 through competition, and alterations in ecosystem functions, and are thus considered among the
469 most destructive feral mammals (Courchamp et al., 2003). Due to their massive ecological impacts,
470 these species have become the target of multiple eradication programs (Courchamp et al., 2003;
471 Barrios-Garcia & Ballari, 2012; Jones et al., 2016). The economic costs found here can also be
472 due to a history of species introduction on islands, for farming, hunting, or biocontrol of previous
473 invasive feral species (Courchamp et al., 2003).

474

475 *Feral vs. wild species*

476 Surprisingly, invasive feral species recorded twice higher average economic costs than wild
477 species. This result suggests that due to the close association between domestic populations and
478 human societies, their impact has likely been more easily observable and quantifiable.
479 Furthermore, it is worth noting that these economic impacts are typically observed in human-
480 structured environments, which are more inclined towards assessing the impacts of domestic
481 species, but also more prone to economic costs (Pimentel et al., 2005). In contrast, the economic
482 impacts on wild species are often overshadowed, as their effects tend to be more ecological in
483 nature rather than directly tied to human economic activities (Pejchar & Mooney, 2009). These
484 ecological impacts encompass changes in biodiversity, ecosystem functioning, and ecological
485 services, which are complex to quantify and may not translate directly into monetary terms (Naeem
486 et al., 2012; Bacher et al., 2018). However, it is also worth highlighting that the sentience and
487 charisma of species matters—independently of their potential impacts—whereby laws around
488 management of invasive feral species (e.g., eradication programs) can encounter opposition from
489 the public and from animal rights groups (Simberloff et al., 2013). A clear example is the case of

490 stray cat colonies in Spain, which are protected by national laws motivated by public perception
491 and animal rights groups, thereby potentially obscuring the scientific evidence around their
492 impacts (Carrete et al., 2022). It should also be highlighted that most domestic species actually
493 belong to different species with respect to their wild counterparts (e.g., *L. glama* and *L. guanicoe*),
494 and accordingly their impacts—when becoming invasive—can be different from those of their
495 wild relatives (Zeder, 2012). These differences can be attributed to the distinct evolutionary paths
496 taken by domestic species, often resulting in varied behavior, ecological needs, and adaptability
497 (Driscoll et al., 2009; Larson & Fuller, 2014). Furthermore, it should be highlighted that this
498 analysis is restricted to mammals and birds, although the inclusion of other species could further
499 magnify the differences found between both groups (i.e., invasive feral and wild species).

500

501 *Conclusions*

502 This study confirms our initial expectations regarding the substantial economic impacts of invasive
503 feral animals with an estimated global cost of at least \$141.95 billion over the past 60 years.
504 However, it is also important to note that some of our predictions were only partially fulfilled due
505 to existing knowledge gaps across taxa and biogeographic regions in the *InvaCost* database
506 (Balzani et al., 2022; Jiang et al., 2022; Soto et al., 2022). Therefore, our results are a conservative
507 estimate of the actual costs, suggesting that the true economic burden may be even higher. Here,
508 we aim to emphasize the possible economic consequences of irresponsible animal husbandry and
509 discourage the acquisition, and particularly the release, of invasive feral species — even those not
510 officially banned. Management of feral populations of domestic animals is fraught with social,
511 ethical, and political complexities. Ownership of these populations is often a subject of debate, but
512 is an issue that is crucial to advance knowledge and direct management. Therefore, we underline
513 the importance of risk assessments including both ecological and economic impact (Soto et al.,
514 2023b) in developing effective laws and tools for managing domesticated species and addressing
515 escaped and stray individuals. Regulations for domesticated pets should focus on controlling feral
516 populations, while risk assessments are needed for livestock and beasts of burden to ensure
517 responsible introductions that minimize impacts on biodiversity and economies. Public awareness
518 and education are crucial, as legislation alone does not guarantee compliance (Patoka et al., 2018).

519

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525

526 **Data availability statement**

527 All data and R codes can be found under <https://github.com/IsmaSA/Domestic-InvaCost-topic>.

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777 **Supplement Note 1.** Description of the descriptors used in the *InvaCost* database

778 *Cost descriptors*

779

780 I. *Method reliability*: This descriptor assesses the reliability of the methods used to estimate
781 the costs of alien species. The costs can be classified as high reliability (e.g., from official
782 reports or peer-reviewed articles with repeatable traceable methods), otherwise classified
783 as low reliability;

784

785 II. *Implementation*: The cost estimates can be classified as either observed or potential.
786 Observed costs refer to incurred costs, while potential costs refer to costs that are predicted
787 or extrapolated over time and/or space, but have not yet been incurred;

788

789 III. *Species*: This descriptor provides the specific invasive alien species to which the incurred
790 costs are attributed;

791

792 IV. *Impacted sector*: This descriptor provides information about which sector or sectors are
793 affected by the presence and spread of alien species, such as agriculture, forestry, or health.
794 In the case of several sectors, these groups were grouped into the category "Diverse";

795

796 V. *Type of cost merged*: This descriptor provides the type of costs (damage, management or
797 mixed) derived from the invasion. Damage costs refer to damages or losses incurred due
798 to the invasion (e.g., costs for damage repair, resource losses or medical care). Management
799 costs refer to investments or expenditure on related activities (e.g., monitoring, prevention,
800 control, eradication). Mixed costs do not distinguish between damage and management
801 costs;

802

803 VI. *Islands*: This descriptor provides information about whether the costs of alien species are
804 present on islands or mainland regions. Those entries recorded as “unspecified” were filled
805 by checking each reference to determine if the costs were produced in islands. Following
806 the classification of *InvaCost*, Australia was considered as an island;

807

808 VII. *Geographic region:* This descriptor provides information about the location(s) where the
 809 recorded costs were incurred;

810

811 VIII. *Official country:* This descriptor provides the specific information about country(s) where
 812 the recorded costs were incurred.

813

814 Description of each category in Impacted sector:

| Sector | Description |
|---------------------------|---|
| Agriculture | Considered at its broadest sense, food and other useful products produced by human activities through using natural and/plant resources from their ecosystems (e.g., crop growing, livestock breeding, beekeeping, land management) |
| Authorities-Stakeholders | Governmental services and/or official organisations (e.g., conservation agencies, forest services, associations) that allocate efforts for the management sensu lato of biological invasions (e.g., control programs, eradication campaigns, research funding) |
| Environment | Impacts on natural resources, ecological processes and/or ecosystem services that have been valued by authors such as disruption of native habitats or degradation of local habitats |
| Fishery | Fish-based activities and services such as fishing and aquaculture |
| Forestry | Forest-based activities and services such as timber production/industries and private forests |
| Health | Every item directly or indirectly related to the sanitary state of people such as vector control, medical care and other derived damage on human productivity and well-being |
| Public and social welfare | Activities, goods or services contributing - directly or indirectly - to the human well-being and safety in our societies, including local infrastructures (e.g., electric system), quality of life (e.g., income, recreational activities), personal goods (e.g., private properties, lands), public services (e.g., transports, water regulation), and market activities (e.g., tourism, trade) |

815

816 **Supplementary Material**

817 Table S1: List of domestic species based on *Domestic Animal Diversity Information System*
 818 database (DAD-IS).

| Scientific name | Common name | Classification |
|-------------------------------|--------------------------|-----------------|
| <i>Anas platyrhynchos</i> | Duck | Livestock |
| <i>Anser anser</i> | Goose | Livestock |
| <i>Axis axis</i> | Chital or axis deer | Livestock |
| <i>Bison bison</i> | American Bison | Beast of burden |
| <i>Bos grunniens</i> | Yak | Beast of burden |
| <i>Bos indicus</i> | Cattle | Livestock |
| <i>Bos taurus</i> | Cattle | Livestock |
| <i>Bos frontalis</i> | Cattle | Livestock |
| <i>Bubalus bubalis</i> | Buffalo | Beast of burden |
| <i>Cairina moschata</i> | Muscovy duck | Livestock |
| <i>Camelus bactrianus</i> | Dromedary Bactrian Camel | Beast of burden |
| <i>Camelus dromedarius</i> | Dromedary | Beast of burden |
| <i>Canis lupus familiaris</i> | Dog | Pet |
| <i>Capra hircus</i> | Goat | Livestock |
| <i>Casuaris casuaris</i> | Cassowary | Livestock |
| <i>Cavia porcellus</i> | Guinea pig | Livestock |
| <i>Cervus albirostris</i> | Thorold's deer | Livestock |

| | | |
|---|-----------------|-----------------|
| <i>Cervus elaphus</i> | Red deer | Livestock |
| <i>Cervus nippon</i> | Sika deer | Livestock |
| <i>Columba livia</i> | Pigeon | Livestock |
| <i>Cyrtonyx montezumae</i> | Montezuma quail | Livestock |
| <i>Dromaius novaehollandiae</i> <i>novaehollandiae</i> | Emu | Livestock |
| <i>Equus asinus</i> | Ass/ Donkey | Beast of burden |
| <i>Equus caballus</i> | Horse | Beast of burden |
| <i>Felis catus</i> | Cat | Pet |
| <i>Gallus gallus</i> | Chicken | Livestock |
| <i>Hydropotes inermis</i> | Water deer | Livestock |
| <i>Lama glama</i> | Llama | Livestock |
| <i>Lama guanicoe</i> | Guanaco | Livestock |
| <i>Lama pacos</i> | Alpaca | Livestock |
| <i>Meleagris gallopavo</i> | Turkey | Livestock |
| <i>Numida meleagris</i> | Guinea fowl | Livestock |
| <i>Oreortyx pictus</i> | Mountain quail | Livestock |
| <i>Oryctolagus cuniculus</i> | Rabbit | Livestock |
| <i>Ovis aries</i> | Sheep | Livestock |
| <i>Pavo cristatus</i> | Peacock | Livestock |
| <i>Pavo muticus</i> | Peacock | Livestock |

| | | |
|-----------------------------------|-----------|-----------|
| <i>Perdix perdix</i> | Partridge | Livestock |
| <i>Phasianus colchicus</i> | Pheasant | Livestock |
| <i>Rangifer tarandus</i> | Deer | Livestock |
| <i>Rhea americana</i> | Nandu | Livestock |
| <i>Struthio camelus australis</i> | Ostrich | Livestock |
| <i>Struthio camelus massaicus</i> | Ostrich | Livestock |
| <i>Struthio camelus syriacus</i> | Ostrich | Livestock |
| <i>Struthio molybdophanes</i> | Ostrich | Livestock |
| <i>Sus</i> sp. | Pig | Livestock |
| <i>Synoicus chinensis</i> | Quail | Livestock |
| <i>Vicugna vicugna</i> | Vicuña | Livestock |