

1 **Benefits do not balance costs of biological invasions**

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42

43 **Abstract**

44 Biological invasions have profound impacts on biodiversity and ecosystem functioning and
45 services, resulting in substantial economic and health costs estimated in the trillions of dollars.
46 Preventing and managing biological invasions are vital for sustainable development, aligning
47 with the goals of the United Nations Biodiversity Conference. However, some invasive species
48 also offer occasional benefits, leading to divergent perceptions among stakeholders and sectors.
49 Claims that invasion science overlooks positive contributions threaten to hinder proper impact
50 assessment and undermine management. Balancing benefits and costs quantitatively is
51 misleading because they coexist without offsetting each other. Any benefits also come at a price,
52 affecting communities and regions differently over time. A considered, integrated approach
53 considering both costs and benefits is necessary for understanding and effective management of
54 biological invasions.

55

56 **Keywords:** invasive species, economic costs, socio-ecological impacts, ethical management.

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59 **1. Introduction**

60 Biological invasions negatively affect biodiversity, ecosystem functioning, and ecosystem
61 services (Simberloff et al. 2013, Pyšek et al. 2020, IPBES 2023). Although not all alien species
62 have reported negative impacts (Bacher et al. 2023), the subset that becomes invasive negatively
63 affects social well-being, reduces cultural diversity, and burdens human well-being and the
64 economy with large costs (Diagne et al. 2021; Stoett et al. 2019). Minimum estimates of the
65 monetary cost of biological invasions are in the order of trillions of dollars (Diagne et al. 2021),
66 comparable to the losses incurred due to natural hazards (Turbelin et al. 2023). Preventing and

67 managing invasions is therefore integral to the sustainable development agenda, as reflected in
68 Target 6 of the Kunming-Montreal Biodiversity Framework Convention on Biological Diversity
69 (2022) and through the recent Intergovernmental Science-Policy Platform on Biodiversity and
70 Ecosystem Services (IPBES) assessment on invasive alien species and their control (IPBES
71 2023).

72 However, some invasive species can occasionally bring benefits to some sectors of society,
73 such as monetary gains (Shackleton et al. 2019a). Thus, actors from different sectors,
74 practitioners, and scientists often have contrasting perceptions about the net sign of their effects
75 on ecosystems or recipient communities and management actions (Jeschke et al. 2014,
76 Shackleton et al. 2022). These perceptions of benefits have led to claims that the field of invasion
77 science focuses exclusively on the negative effects of invasive species, and overlooks positive
78 contributions to economies and ecosystems (Boltovskoy et al. 2022, Sax et al. 2022, 2023).
79 Although these arguments are flawed because they conflate ‘alien’ with ‘invasive alien species’,
80 the fact that some invasive species have benefits is undisputed (Vimercati et al. 2020,
81 Kourantidou et al. 2022, IPBES 2023). Consequently, considering the occurrence and
82 relationships between costs and benefits is necessary to contextualise the management of
83 invasive species. In this regard, we explain that this must be done with caution because a direct,
84 quantitative balance of benefits and costs is overly simplistic and misleading for three main
85 reasons we describe in detail in the following sections of this forum paper: (i) benefits of
86 invasive species as a collective have never been demonstrated to be as high as their massive
87 documented costs, (ii) benefits do not offset costs (they only exist in parallel to costs), and (iii)
88 benefits always come at a price, because they are context-dependent and affect different
89 stakeholders or regions at different times. Emphasising economic benefits over negative impacts,
90 drawing from ambiguities, and uncertain or unpredictable effects can be risky to conservation

91 goals, because it can bias and hinder the proper assessment of all impacts, ultimately
92 undermining the management of biological invasions.

93

94 **2. Benefits are rarer than costs**

95 Invasive species are plants, animals, pathogens, and other organisms that have evolved outside of
96 a recipient ecosystem, and that can cause economic or environmental harm or adversely affect
97 human health (Convention on Biological Diversity 2021, IPBES 2023). Importantly, non-native
98 species or populations that are farmed or cultivated are generally not included in this category
99 when they do not spread outside human-controlled environments. This means that the economic
100 benefits of agriculture, aquaculture, and forestry are, for the most part, unrelated to biological
101 invasions and therefore, many of these benefits are extraneous and irrelevant to the invasions
102 discourse. In addition, because invasive species are associated with negative impacts, they
103 inherently imply the presence of costs, whether as a direct result of their negative consequences
104 on ecosystems, or indirectly through expenditure on their control. Documented benefits of
105 invasive species are therefore typically by-products. In fact, any benefits are exceptions or
106 special cases in the face of the massive ecological impacts of invasive species (Simberloff et al.
107 2013, IPBES 2023), or the result of management seeking their control or removal. While
108 systematic comparisons of costs and benefits are sparse, a recent systematic review analysing the
109 number of cases of costs and benefits (labelled as 'undesirable' and 'desirable' effects,
110 respectively) of biological invasions confirmed the expectation that the presence of costs is more
111 frequent and affects more sectors than benefits (Kelsch et al. 2020, IPBES 2023) (Fig. 1).
112 Similarly, the recent IPBES assessment on invasive species highlights that reports from some
113 Indigenous Peoples and local communities documented that 92% of the impacts on nature from
114 invasive alien species were negative, with only 8% being positive (IPBES 2023). More broadly,

115 of 3783 documented impacts of invasive alien species on quality of life in the IPBES report,
116 more than 85% are negative and far fewer (15%) are positive for good quality of life (Bacher et
117 al. 2023). In addition, the overall benefits of invasive species have never been demonstrated to be
118 anything other than small compared to quantified costs. The more than US\$2 trillion (2017
119 currency value) costs of biological invasions already recorded in the (still expanding; Ahmed et
120 al. 2023) *InvaCost* database represent a massive, and yet conservative estimate, which could
121 conceal a much higher true cost (Diagne et al. 2021, Leroy et al. 2021, Ahmed et al. 2023).

122 Moreover, benefits related to financial gains can often be easily quantifiable for some
123 stakeholders, such as those in fisheries and tourism (Kerr 2019), but the costs are typically not as
124 straightforward to assess and quantify. These difficulties are particularly large when pertaining to
125 the ecological damages to recipient ecosystems (IPBES 2023). Figure 1 shows that the greatest
126 costs of biological invasions are those that affect the environment, while most of their economic
127 benefits are present in local communities (e.g., businesses), fisheries, and livestock farming —
128 more visible, anthropocentric, and quantifiable sectors.

129 The temporal scale at which benefits and costs are generated is also necessary for comparison.
130 The magnitude and range of social and environmental impacts (and potential associated
131 economic costs) typically manifest over longer durations than any associated economic benefits,
132 thereby presenting an immense challenge to management and policy. Indeed, ecological effects
133 are often undetectable when invasive populations are first introduced (Daly et al. 2023), and
134 while evidence for monetary impacts alone should not underpin conservation actions, assigning
135 monetary values to their negative impacts can be challenging or impossible in these early stages.
136 For example, the short-term economic gains from commercial fisheries targeting the Nile perch
137 (*Lates niloticus*) in Lake Victoria have ultimately come at the vastly larger, long-term expense of
138 the ecological and socio-economic integrity of large lake areas, driving one of the greatest

139 extinction events of hundreds of native and endemic fish species in modern history (Aloo et al.
140 2017).

141 Finally, while monetary metrics allow for the quantification of some benefits and costs, the
142 effects of biological invasions are often complex and difficult to value in monetary terms, with
143 invasions sometimes benefitting certain taxa or ecosystem services while concurrently affecting
144 others negatively.

145

146 **3. Benefits do not negate costs**

147 Even when generated by the same invasive population, the costs *versus* benefits of invasive
148 species should not be compared to the expenditures *versus* revenue in a simple accounting
149 framework (IPBES 2023), particularly because they usually target different sectors (Figure 1)
150 that can be affected over different periods. Benefits can stem from (i) exploiting the invasive
151 species directly, and (ii) profitable activities indirectly leading to an introduction. In the first
152 case, intentional introductions of species such as the red swamp crayfish (*Procambarus clarkii*)
153 in Europe and Atlantic salmon (*Salmo salar*) in Chilean Patagonia for aquaculture have indeed
154 brought economic benefits for the companies commercially exploiting them (Souty-Grosset et al.
155 2016, Figueroa-Muñoz et al. 2022). However, these species have had substantial negative
156 impacts in the wild across multiple sectors, including costs to agriculture, water management,
157 and fishing livelihoods. In the second case of trade-driven accidental introductions, examples
158 include profitable timber imports or agricultural products that can precipitate unintended insect
159 invasions. There can also be unintentional escapes of species used from fur farming, insect
160 farming, ornamental horticulture, or as pets (Hulme et al. 2021, Carpio et al. 2020, Bang and
161 Courchamp 2021). Therefore, the resultant economic costs are generally borne by other
162 industries than those responsible for the introductions or end-consumers who are faced with

163 higher prices, which account for the invasion externality of yield losses, damage to
164 infrastructure, or other types of costs (Diagne et al. 2021, Kourantidou et al. 2022). When
165 economic profits are generated by the invasive species, costs of invasions continue to exist even
166 if implicit.

167

168 **4. Benefits to some are costly to others**

169 Despite biological invasions being one of the five major direct drivers of biodiversity loss (e.g.,
170 IPBES 2019, 2023) and damage costs analogous to those of natural disasters (Turbelin et al.
171 2023), the emphasis on benefits, either ecosystem, social or monetary, makes them an oddity in
172 the fields of global change research. As a parallel, it is possible to think of the economic benefits
173 of climate change (e.g., for sectors involved in climate-change adaptation), of habitat destruction
174 (e.g., for real-estate developers and, more generally, any industry benefiting from urban
175 populations or agriculture), and of overexploitation (e.g., for recreation or commerce). It is
176 doubtful that researchers studying these benefits would assign biases to the collation of costs, or
177 present global change as desirable because of such restricted benefits. Such is, however, the
178 status quo in critiques of invasion science. An undue emphasis on the benefits of biological
179 invasions can also prevent or delay their management, ultimately leading to higher long-term
180 economic costs and negative ecological and health impacts (Leung et al. 2002). Indeed, recent
181 estimates for delayed management, even of a single species, are tens of millions of dollars per
182 year (Ahmed et al. 2022).

183 In a world where economic growth is still prioritised over nature conservation, promoting
184 such sporadic and short-term benefits might also create dependencies within affected
185 communities, undermining management and ultimately disrupting ecosystem structure and
186 function (Vitule and Pelicice 2023). For example, the development of the charcoal industry and

187 bio-power plants around invasive species like the shrub mesquite (*Prosopis juliflora*) was
188 intended to improve livelihoods, generate energy, and manage the species sustainably in low- to
189 middle-income Asian and African nations (Mwangi and Swallow 2008, Walter and Armstrong
190 2014). Nevertheless, the introduced shrub outcompeted native plants and disrupted an entire
191 ecosystem, threatening local community resources and cultural values. The perceptions can be
192 also influenced by other context dependencies (Kourantidou et al. 2022) e.g., previous land use
193 and features of landscapes where invasive trees in barren grasslands (*Acacia dealbata* in South
194 Africa) are perceived as valuable assets for soil erosion regulation and resource provision, but
195 cause damage to households, reduce crop production, and serve as hideouts for criminals
196 (Ngorima and Shackleton 2019). However, treating an invasive species as a resource is often not
197 the preference of local communities (IPBES 2023) and has shifted the focus to perpetuating the
198 benefit instead of eradicating the problem, often leading to privatisation of the benefits and
199 socialisation of the costs.

200

201 **5. Ethical management must integrate all positive and negative effects**

202 Despite the three major shortcomings detailed above, managing biological invasions does require
203 accounting for any potentially positive effects and the implications to community members who
204 will be affected one way or another. In general, non-monetary evaluation methods such as
205 qualitative assessments, social impact analyses, and environmental impact assessments could be
206 used to infer environmental and other costs across invaded systems. This can be done by
207 combining semi-quantitative approaches integrating the positive and beneficial effects of
208 biological invasions in existing frameworks (Shackleton et al. 2019a, Vimercati et al. 2020),
209 such as the *Environmental Impact Classification for Alien Taxa* (IUCN 2020, EICAT+;
210 Vimercati et al. 2022) and the *Socio-Economic Impact Classification for Alien Taxa* (Bacher et

211 al. 2018), with quantitative approaches through databases like *InvaCost* (Diagne et al. 2020;
212 Ahmed et al. 2023).

213 Decision-makers should strive to communicate the risks and impacts, negative and positive,
214 of biological invasions across different sectors of the communities directly affected, and involve
215 affected community members in decision-making processes (Reed et al. 2023, IPBES 2023).
216 Equally, they should identify and raise awareness of alternative native resources that could
217 provide similar benefits. The *Nature's Contributions to People* framework (Díaz et al. 2018)
218 acknowledges that nature has a plurality of values, including intrinsic, instrumental, and
219 relational values, and that decision-making and policy design should be informed by these
220 different and complementary perspectives (IPBES 2019). For example, Indigenous Peoples and
221 local communities might perceive the introduction of a non-native fish as an opportunity for
222 fishing (Lima et al. 2010), or a non-native tree species as a source of shade and timber (Kull et
223 al. 2019) without being immediately aware of the negative impacts on native species
224 populations, on loss of arable land or health issues (Shackleton et al. 2019b). Differences in
225 perspectives between regions can be due to differences in socio-economic development and
226 access to natural resources (Meyer and Fourdrigniez 2019). That the public is sometimes more
227 aware of direct, positive effects than indirect, negative ones (Sax et al. 2022) is not a good reason
228 to abandon the management of biological invasions.

229

230 **Conclusion**

231 Positive effects of biological invasions do occasionally exist and should not be ignored by
232 policymakers (Shackleton et al. 2019a, Vimercati et al. 2020). However, acknowledging these
233 benefits does not negate the necessity of evaluating the overall impact (IPBES 2023). Moreover,
234 global efforts to create evidence for the benefits of biological invasions are lacking, and do not

235 match the recently collated evidence for its costs. In the absence of such matching evidence, it is
236 incorrect to propose a comparability of the benefits and costs of biological invasions. While
237 benefits, unlike costs, lack a monetary synthesis, it is important to recognise that benefits and
238 costs often affect different stakeholders, operate over divergent time scales, and are viewed
239 through diverse socio-ecological perspectives. Promoting the benefits of biological invasions can
240 also hinder management and thus increase costs in the long term. The evidence and arguments
241 we provide here demonstrate that it is ethically and scientifically dubious to argue for prioritising
242 the limited economic benefits of biological invasions in the face of their overwhelming negative
243 ecological, economic, and social impacts. Calls for shifting focus to the monetary benefits of
244 biological invasions risks undermining the much-needed awareness and support to mitigate them.

245

246 **References**

- 247 Ahmed DA, et al. 2022. Managing biological invasions: the cost of inaction. *Biological*
248 *Invasions* 24:1927-1946.
- 249 Ahmed DA, et al. 2023. Recent advances in availability and synthesis of the economic costs of
250 biological invasions. *BioScience* 73:560-574.
- 251 Aloo PA, Njiru J, Balirwa JS, Nyamweya CS. 2017. Impacts of Nile perch, *Lates niloticus*,
252 introduction on the ecology, economy and conservation of Lake Victoria, East Africa.
253 *Lakes and Reservoirs: Science, Policy and Management for Sustainable Use* 22:320-333.
- 254 Bacher S, et al. 2018. Socio-economic impact classification of alien taxa (SEICAT). *Methods in*
255 *Ecology and Evolution* 9:159-168.
- 256 Bacher, S., 2023. Chapter 4: Impacts of invasive alien species on nature, nature's contributions to
257 people, and good quality of life. In: Thematic Assessment Report on Invasive Alien
258 Species and their Control of the Intergovernmental Science-Policy Platform on
259 Biodiversity and Ecosystem Services. Roy, H. E., Pauchard, A., Stoett, P., and Renard
260 Truong, T. (eds.). IPBES secretariat, Bonn, Germany. doi:10.5281/zenodo.7430731
- 261 Bang A, Courchamp F. 2021. Industrial rearing of edible insects could be a major source of new
262 biological invasions. *Ecology Letters* 24:393-397.

- 263 Boltovskoy D, Guiaşu R, Burlakova L, Karatayev A, Schlaepfer MA, Correa N. 2022.
264 Misleading estimates of economic impacts of biological invasions: Including the costs
265 but not the benefits. *Ambio* 51:1786-1799.
- 266 Carpio AJ, Álvarez Y, Oteros J, León F, Tortosa FS. 2020. Intentional introduction pathways of
267 alien birds and mammals in Latin America. *Global Ecology and Conservation* 22:e00949.
- 268 Convention on Biological Diversity. 2021. What are invasive alien species? (20 July
269 2023; cbd.int/idb/2009/about/what).
- 270 Convention on Biological Diversity. 2022. Target 6 of the Kunming-Montreal
271 Biodiversity Framework. (07 January 2024; cbd.int/gbf)
- 272 Daly EZ, et al. 2023. A synthesis of biological invasion hypotheses associated with the
273 introduction–naturalisation–invasion continuum. *Oikos* 2023:e09645
- 274 Diagne C, Catford JA, Essl F, Nuñez MA, Courchamp F. 2020. What are the economic costs of
275 biological invasions? A complex topic requiring international and interdisciplinary
276 expertise. *NeoBiota* 63:25-37.
- 277 Diagne C, Leroy B, Vaissière AC, Gozlan RE, Roiz D, Jarić I, Salles JM, Bradshaw CJA,
278 Courchamp F. 2021. High and rising economic costs of biological invasions worldwide.
279 *Nature* 592:571-576.
- 280 Díaz S, et al. 2018. Assessing nature's contributions to people. *Science* 359:270-272.
- 281 Figueroa-Muñoz G, Correa-Araneda F, Cid-Aguayo B, Henríquez A, Arias L, Arismendi I,
282 Gomez-Uchida D. 2022. Co-management of Chile's escaped farmed salmon. *Science*
283 378:1060-1061.
- 284 Hulme PE. 2021. Unwelcome exchange: International trade as a direct and indirect driver of
285 biological invasions worldwide. *One Earth* 4(5):666-679.
- 286 IPBES 2019. Global Assessment Report on Biodiversity and Ecosystem Services of the
287 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services.
288 Brondizio ES, Settele J, Díaz S, Ngo HT (eds.). IPBES Secretariat, Bonn, Germany. 1148
289 pages. doi:10.5281/zenodo.3831673
- 290 IPBES 2023. Summary for Policymakers of the Thematic Assessment Report on Invasive Alien
291 Species and their Control of the Intergovernmental Science-Policy Platform on
292 Biodiversity and Ecosystem Services. Roy HE, et al. (eds.). IPBES Secretariat, Bonn,
293 Germany. doi:10.5281/zenodo.7430692.
- 294 IUCN 2020. IUCN EICAT Categories and Criteria. The Environmental Impact Classification for
295 Alien Taxa (EICAT). IUCN Gland, Switzerland and Cambridge, UK.
- 296 Jarić I, et al. 2020. The role of species charisma in biological invasions. *Frontiers in Ecology and
297 the Environment*, 18:345-353.

- 298 Jeschke JM, et al. 2014. Defining the impact of non-native species. *Conservation Biology*
299 28:1188-1194.
- 300 Kelsch A, Takahashi Y, Dasgupta R, Mader AD, Johnson BA, Kumar P. 2020. Invasive alien
301 species and local communities in socio-ecological production landscapes and seascapes: a
302 systematic review and analysis. *Environmental Science and Policy* 112:275-281.
- 303 Kerr G. 2019. Himalayan tahr (*Hemitragus jemlahicus*) recreational hunting values. *Wildlife*
304 *Research* 46:114-126.
- 305 Kourantidou M, Haubrock PJ, Cuthbert RN, Bodey TW, Lenzner B, Gozlan RE, Nuñez MA,
306 Salles JM, Diagne C, Courchamp F. 2022. Invasive alien species as simultaneous benefits
307 and burdens: trends, stakeholder perceptions and management. *Biological Invasions*
308 24:1905-1926.
- 309 Kull CA, Harimanana SL, Radaniela Andrianoro A, Rajoelison LG. 2019. Divergent perceptions
310 of the ‘neo-Australian’ forests of lowland eastern Madagascar: invasions, transitions, and
311 livelihoods. *Journal of Environmental Management* 229:48-56.
- 312 Leroy B, et al. 2021. Global Costs of Biological Invasions: Living Figure. (08 January 2024;
313 borisleroy.com/invacost/invacost_livingfigure.html)
- 314 Leung B, Lodge DM, Finnoff D, Shogren JF, Lewis MA, Lamberti G. 2002. An ounce of
315 prevention or a pound of cure: bioeconomic risk analysis of invasive species. *Proceedings*
316 *of the Royal Society B: Biological Sciences* 269:2407-2413.
- 317 Lima FP, Latini AO, De Marco Júnior P. 2010. How are the lakes? Environmental perception by
318 fishermen and alien fish dispersal in Brazilian tropical lakes. *Interciencia* 35:84-91.
- 319 Meyer J-Y, Fourdrigniez M. 2019. Islander perceptions of invasive alien species: the role
320 of socio-economy and culture in small isolated islands of French Polynesia (South
321 Pacific). *Island Invasives: Scaling Up to Meet the Challenge*. In Veitch CR, Clout MN,
322 Martin AR, Russell, JC, West CJ (eds.), pp. 510-516. IUCN Occasional Paper SSC no.
323 62. Gland, Switzerland.
- 324 Mwangi E, Swallow B. 2008. *Prosopis juliflora* invasion and rural livelihoods in the Lake
325 Baringo area of Kenya. *Conservation and Society* 6: 130–140.
- 326 Ngorima, A., Shackleton, C. M. 2019. Livelihood benefits and costs from an invasive alien tree
327 (*Acacia dealbata*) to rural communities in the Eastern Cape, South Africa. *Journal of*
328 *Environmental Management* 229, 158-165.
- 329 Pyšek P, et al. 2020. Scientists’ warning on invasive alien species. *Biological Reviews* 95:1511-
330 1534.
- 331 Reed EMX, Schenk T, Brown BL, Rogers H, Haak DC, Drake JC, Barney JN. 2023. Holistic
332 valuation of non-native species requires broadening the tent. *Trends in Ecology and*
333 *Evolution* 38:497-498.

334 Sax DF, Schlaepfer MA, Olden JD. 2022. Valuing the contributions of non-native species to
335 people and nature. *Trends in Ecology and Evolution* 37:1058-1066.

336 Sax DF, Schlaepfer MA, Olden JD. 2023. Identifying key points of disagreement in non-native
337 impacts and valuations. *Trends in Ecology and Evolution* 38:501-504.

338 Shackleton RT, et al. 2019a. Explaining people’s perceptions of invasive alien species: A
339 conceptual framework. *Journal of Environmental Management* 229 10-26.

340 Shackleton RT, Shackleton CM, Kull CA. 2019b. The role of invasive alien species in shaping
341 local livelihoods and human well-being: a review. *Journal of Environmental Management*
342 229:145–157.

343 Shackleton RT, Vimercati G, Probert AF, Bacher S, Kull CA, Novoa A. (2022).
344 Consensus and controversy in the discipline of invasion science. *Conservation Biology*,
345 36(5), e13931.

346 Simberloff D, et al. 2013. Impacts of biological invasions: what’s what and the way forward.
347 *Trends in Ecology and Evolution* 28:58-66.

348 Souty-Grosset C, Anastácio PM, Aquiloni L, Banha F, Choquer J, Chucholl C, Tricarico E.
349 2016. The red swamp crayfish *Procambarus clarkii* in Europe: Impacts on aquatic
350 ecosystems and human well-being. *Limnologica* 58:78-93.

351 Stoett P, Roy HE, Pauchard A. 2019. Invasive alien species and planetary and global health
352 policy. *The Lancet Planetary Health* 3:e400-e401.

353 Turbelin AJ, Cuthbert RN, Essl F, Haubrock PJ, Ricciardi A, Courchamp F. 2023. Biological
354 invasions are as costly as natural hazards. *Perspectives in Ecology and Conservation*
355 21:143–150.

356 Vimercati G, Kumschick S, Probert AF, Volery L, Bacher S. 2020. The importance of assessing
357 positive and beneficial impacts of alien species. *NeoBiota* 62:525.

358 Vimercati G, et al. 2022. The EICAT+ framework enables classification of positive impacts of
359 alien taxa on native biodiversity. *PLoS Biology* 20:e3001729.

360 Vitule JRS, Pelicice FM. 2023. Care needed when evaluating the contributions of non-native
361 species. *Trends in Ecology and Evolution* 38:499-500.

362 Walter KJ, Armstrong KV. 2014. Benefits, threats and potential of *Prosopis* in South India.
363 *Forests, Trees and Livelihoods* 23:232-247.

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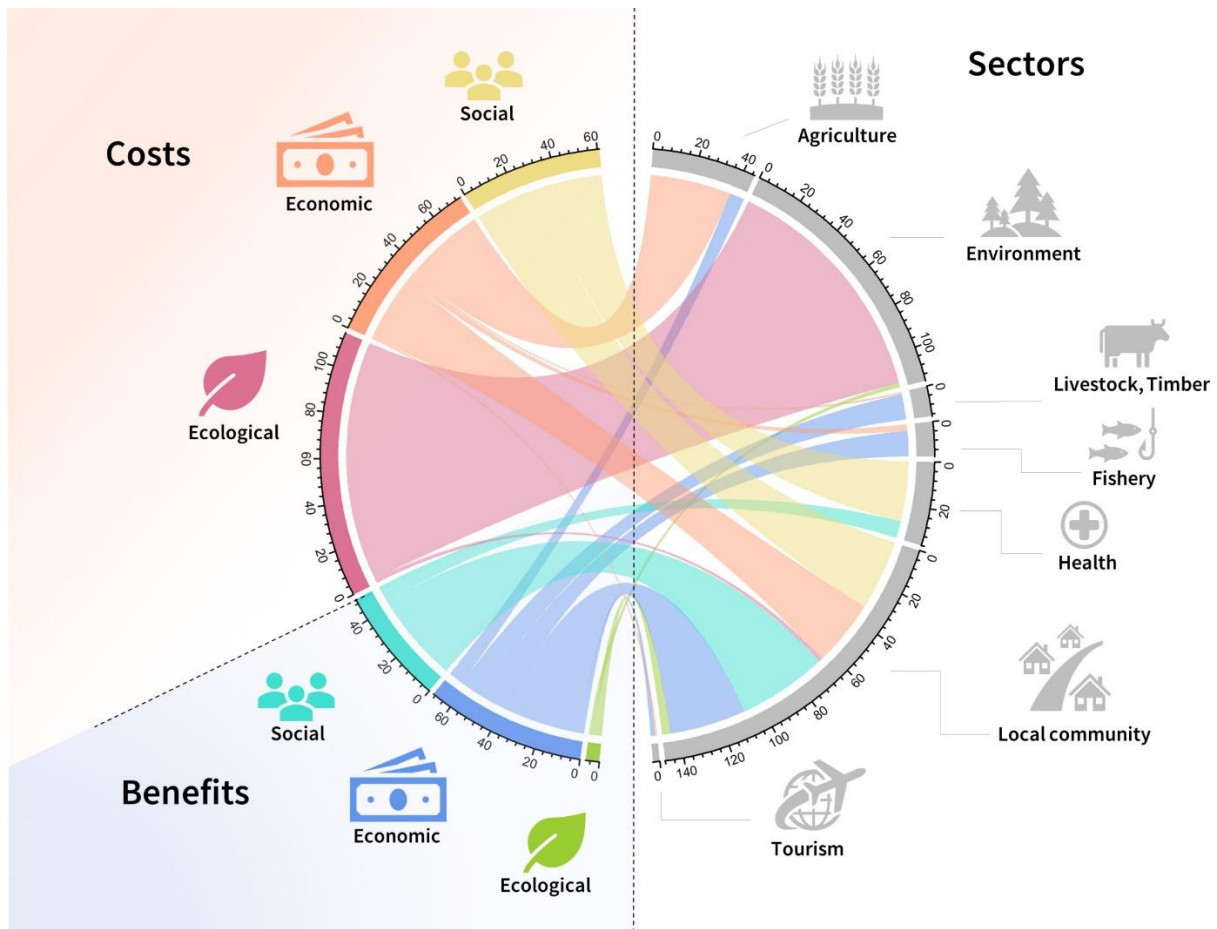
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374

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380 **Fig. 1. Number of cases of costs and benefits of invasive species as indicated by respondents**
 381 **from local communities and recorded in the database of Kelsch et al. (2020), and flows**
 382 **towards the sector(s) they affect.** We have labelled the 'desirable' positive effects as beneficial,
 383 while those effects described as 'undesirable' as costs. There are more cases of costs than
 384 benefits, and different sectors are implicated. The database, which separates the desirable and
 385 undesirable effects of invasive species, was reclassified as benefits and costs (economic, social,
 386 ecological) related to sectors (local community, health, tourism, fishery, farmers, environment,
 387 agriculture). Social costs are non-monetisable, such as impediments to movement, health
 388 problems, and cultural practices. Ecological costs include biodiversity loss and environmental

389 degradation. Economic costs include loss of arable lands, production (e.g., crops, fisheries,
390 livestock), primary resources, and water supply. Social benefits include cultural practices,
391 resource provisioning (e.g., food and fuel), and medicinal uses. Economic benefits include
392 commercial use of resources (e.g., fuel, fishery, timber), and tourism. Ecological benefits are
393 reported for increasing water supply, soil fertility, and control of other invasive species, although
394 they are secondary to previous impacts reported (e.g., drought, ecosystem degradation,
395 introduction of invasive species).